RECEIVED

06 JAN 17 AM 9:31

HIGH PRODUCTION VOLUME (HPV) CHEMICAL CHALLENGE PROGRAM

TEST PLAN For:

Diisopropyl Ether (DIPE) CAS No. 108-20-3

Prepared by:

ExxonMobil Chemical Company Shell Chemical LP

For:

American Chemistry Council, Isopropanoi Panel, Diisopropyl Ether HPV Task Group

December 12, 2005

EXECUTIVE SUMMARY

Under the U.S. Environmental Protection Agency (EPA) High Production Volume (HPV) Chemical Challenge Program (Program), ExxonMobil Chemical Company and Shell Chemical LP committed thru the Isopropanol Panel, DIPE Task Group of the American Chemistry Council (ACC) to voluntarily compile a Screening Information Data Set (SIDS) that can be used for an initial hazard assessment of diisopropyl ether (DIPE), CAS No. 108-20-3. The data described in this test plan should be used for the purposes of HPV Program and not for regulatory cleanup or criteria development processes.

The assessment includes data for physicochemical, environmental fate, and mammalian and environmental effect endpoints included in the U.S. HPV Program. Additional mammalian data beyond the SIDS endpoints are supplied with this submission.

A literature search identified data for all endpoints under the U.S. EPA HPV Program for DIPE (Table 6); however, the studies for two environmental toxicity endpoints were not sufficiently reliable. All the mammalian toxicity studies in this assessment were assessed as adequate, with reliable data that suggest DIPE generally presents a low order of toxicity for human health. Although data were identified for all aquatic toxicity endpoints, the acute invertebrate and alga toxicity endpoints did not have measured data that were equivalent in quality to the data used to characterize the fish acute toxicity endpoint. Therefore, studies will be conducted to develop measured data that characterize DIPE for these endpoints. The available aquatic toxicity data suggest that DIPE presents a low order of toxicity for the environment.

Results of Mackay Level I distribution modeling at steady state show that DIPE will partition primarily to the air compartment (97.8%), with a negligible amount partitioning to water (2.1%) and soil (0.1%). Level III modeling indicates that at steady state, water is the primary compartment on a percentage basis when the default emission to this compartment is included in the calculations. However, Level III modeling may not be representative of the ultimate disposition of DIPE because default emissions, which use 1000 kg/h/compartment, are not representative of chemical discharge, including to groundwater.

DIPE is volatile, and volatilized DIPE will be quickly degraded in the atmosphere via indirect photodegradation. The DIPE half-life from hydroxyl radical attack is calculated to be approximately 5 hours. Aqueous photolysis and hydrolysis will not contribute to the transformation of DIPE in aquatic environments because it is either poorly or not susceptible to these reactions.

DIPE has a low potential to biodegrade based on results of biodegradation testing. Bioaccumulation of DIPE is unlikely, based on a low bioconcentration factor (bioconcentration factor = 2.95).

Results of testing suggest that DIPE exhibits low aquatic toxicity, based on an analytically measured fish acute toxicity value of 92 mg/L (96-hour LC₅₀) and a nominally measured acute invertebrate toxicity value reported as 190 mg/L (48-hour EC₅₀). Two additional fish acute toxicity studies are available with results that range up to 900 mg/L for a 24- and 96-hour study. Although there are no reliable measured data

for an alga, calculated data are available. Calculated 96-hour EC₅₀ and chronic values are 135 and 10 mg/L, respectively.

Available data shows DIPE to be or a low order of acute oral, dermal, and inhalation toxicity with LD $_{50}$ values in excess of 2000 mg/kg and an inhalation LC $_{50}$ >20 mg/kg. High concentrations of DIPE cause CNS depression which is readily reversible on cessation of exposure. DIPE is not a skin irritant but prolonged/repeated contact may cause defatting of the skin, which can lead to dermatitis. The vapours and liquid may be irritating to the eyes at 800 ppm, but not at 500 ppm

There are no reports of human systemic toxicity associated with acute DIPE exposure. DIPE is not expected to be a skin sensitizer. Increased liver weights in rats without histopathology may have been an adaptive effect. There were no adverse effects in repeat dose animal studies other than the male rat kidney effect, which may not be relevant to humans. DIPE does not appear to be a primary reproductive or developmental toxicant. DIPE is not considered to be neurotoxic. DIPE is not genotoxic *in vitro* and is not considered a mutagenic or carcinogenic hazard.

Existing DIPE data for the HPV Program are summarized below. Two studies will be conducted, acute invertebrate and alga toxicity, to provide a consistent and reliable analytically measured data set for the aquatic endpoints.

DIPE Data Availability and Adequacy for Endpoints in the HPV Program

Endpoint	Data Availability	Reference	First Author/ Source	Date
Physical/Chemical Properties	S III.			
2.1 Melting Point	Adequate measured	22		
2.2 Boiling Point	Adequate measured	22	Lide et al.	1997
2.3 Density	Adequate measured	22		
2.4 Vapor Pressure	Adequate measured	6	Daubert and Danner	1989
2.5 Partition Coefficient	Adequate	15	Hansch et al.	1995
2.5 Partition Coefficient	measured	8	Eadsforth	1983
2.6 Water Solubility	Adequate measured	12	Gerhartz et al.	1987

Endpoint	Data Availability	Reference	First Author/ Source	Date
Environmental Fate				
	Technical	16	Harris	1982a
3.1 Photodegradation	discussion (direct)	35	Zepp and Cline	1977
(direct and indirect)	Computer model (indirect)	9	EPI Suite [™]	2000
3.2 Stability in Water	Technical	13	Gould	1959
	discussion	17	Harris	1982b
3.3 Transport between Environmental	Computer	24	Mackay	1998
Compartments (Level I and III)	model	25	Mackay et al.	1996
3.4 Biodegradation	Adequate measured	30	Stone and Watkinson	1983
3.5 Bioaccumulation	Computer model	9	EPI Suite™	2000
Environmental Toxicity				
	Adequate measured	3	Broderius and Kahl	1985
	Adequate measured	11	Geiger et al.	1986
4.1 Acute/prolonged toxicity to	Adequate measured	33	Veith et al.	1983
Fish	Computer model	10	ECOSAR	2004
	Measured (not reliable)	2	Bridie <i>et al.</i>	1979
	Measured (not reliable)	7	Dawson	1975/77
4.2 Acute Toxicity to Aquatic	Measured (nominal)	29	Stephenson	1983
Invertebrates	Computer model	10	ECOSAR	2004
A O Taviait A A A Service Disease	Measured (not reliable)	29	Stephenson	1983
4.3 Toxicity to Aquatic Plants	Computer model	10	ECOSAR	2004

	ndpoint	Data Availability	Reference	First Author/ Source	Date
Mammalian '	Toxicity				
5.1.1. Acute (Oral Toxicity	Adequate	20	Kimura et al.	1971
0.1.1.7.0dto (measured	23		
5.1.2 Acute Ir	nhalation Toxicity	Adequate measured	23	Machle et al.	1939
5.1.3 Acute Dermal Toxicity		Adequate measured	23		
5.2 Sensitizat	ion	Adequate measured	34	Wass and Belin	1990
5.2 Panastad	Dose Toxicity	Adequate measured	5	Dalbey and Feuston	1996
3.3 Nepeateo	Dose Toxicity		27	Rodriguez and Dalbey	1997
5.4 Genetic Toxicity	Bacterial mutagenicity	Adequate	4	Brooks <i>et al</i> .	1998
In Vitro Chromosomal damage		measured	•	Brooke of un.	1000
5.5 Developmental Toxicity/Teratogenicity		Adequate measured	5	Dalbey and Feuston	1996
5.6 Additional Information -		Adequate	18	Hine et al.	1955
Sensory Humans	Irritation in	measured	28	Silverman <i>et</i> al.	1946

TABLE OF CONTENTS

	Page
EXECUTIVE SUMMARY	2
I. INTRODUCTION	7
II. CHEMICAL DESCRIPTION, MANUFACTURING PROCESS, USE, AN EXPOSURE	ND 7
III. TEST PLAN RATIONALE AND DATA SUMMARY	8
A. Physicochemical Data Table 1. Selected Physico-Chemical Properties for DIPE.	8 9
B. Environmental Fate Data	9
Table 2. Environmental distribution as calculated by the Mackay (1998) Level I fugacity model.	11
Table 3. Environmental distribution as calculated by the Mackay (1998) Level III fugacity model.	12
C. Aquatic Toxicity Data	12
Table 4. Aquatic Toxicity Studies for DIPE.	13
D. <u>Human Health Effects Data</u>	13
Table 5. Mammalian Toxicity Studies for DIPE.	17
IV. TEST PLAN SUMMARY Table 6. DIRE Date Availability and Adaptive for Endociate in the	18
Table 6. DIPE Data Availability and Adequacy for Endpoints in the HPV Program.	18
V. REFERENCES	21
APPENDIX Data Dossier (Robust Study Summaries) for Diisopropyl Eth (CAS No. 108-20-3)	ner 24

TEST PLAN FOR DIISOPROPYL ETHER (CAS No. 108-20-3)

I. INTRODUCTION

Under the U.S. Environmental Protection Agency (EPA) High Production Volume (HPV) Chemical Challenge Program (Program), ExxonMobil Chemical Company and Shell Chemical LP committed thru the Isopropanol Panel, DIPE Task Group of the American Chemistry Council (ACC) to voluntarily compile a Screening Information Data Set (SIDS) that can be used for an initial hazard assessment of diisopropyl ether (DIPE), CAS No. 108-20-3. The data described in this test plan should be used for the purposes of HPV Program and not for regulatory cleanup or criteria development processes.

The assessment includes data for selected physicochemical, environmental fate, and mammalian and environmental effect endpoints required by the U.S. HPV Program. Additional data beyond the SIDS endpoints (irritation, sensitization, and neurotoxicity) are supplied with this submission.

Procedures to assess the reliability of selected data for inclusion in this test plan were based on guidelines described by Klimisch *et al.* (1997) and identified within the U.S. EPA (1999a) document titled *Determining the Adequacy of Existing Data*. The following sections describe DIPE and its manufacturing process, and data used to characterize the various endpoints in the HPV Program. Additional mammalian data beyond SIDS requirements are also provided together with exposure information that allow for further assessment of data completeness. After a review of the existing data, the sponsors believe that reliable and consistent data have been identified that can characterize all but two SIDS endpoints. Two studies will be conducted, acute invertebrate and alga toxicity, to provide a consistent and reliable measured data set for the environmental endpoints.

II. <u>CHEMICAL DESCRIPTION, MANUFACTURING PROCESS, USE, AND EXPOSURE</u>

DIPE is a small molecular weight ether represented with the following chemical formula and structures:

$$C_6H_{14}O$$
 (CH₃)₂CHOCH(CH₃)₂ H_3C CH₃ C(H) - O - C(H) H_3C CH₃

DIPE is manufactured by a series of chemical reactions as a co-product in the synthesis of isopropyl alcohol (IPA), typically including a propane feedstock and water to produce a high purity product. In a first reaction zone the propane in a feedstock (after removal of hydrocarbons containing four or more carbon atoms from the feedstock via fractionation) is dehydrogenated in the presence of a dehydrogenation catalyst to form propylene. After removing hydrogen, the propane and propylene mixture generated in the first reaction zone is separated into propane-enriched and propylene-enriched streams. The propylene-enriched stream contains at least 65% (wt.) propylene. The propane-enriched stream is recycled to the feedstock fractionation unit, and the propylene in the propylene-enriched stream is reacted with water in a second reaction zone in the presence of an acidic catalyst to form IPA; some of this can concurrently react with propylene to produce diisopropyl ether. A portion of the second reaction zone

effluent is recycled to the second reaction zone, and the remainder may be collected or further separated to provide a high purity disopropyl ether product.

DIPE can be used as a gasoline blending component. Other uses are as a solvent for animal, vegetable, and mineral oils. It is also used as an extraction solvent for dewaxing of paraffin-based oil products. DIPE can be used as a solvent in medicine production and paint cleaning.

Because DIPE is volatile, inhalation is expected to be the major route of human exposure and ingestion a minor route. While the HPV Program is a hazard data collection effort and does not require evaluation of potential environmental impacts to groundwater, this may be a consideration for DIPE within a risk assessment framework. DIPE has been found in groundwater and drinking water wells within the U.S. Odor and taste studies indicate that DIPE has a very low odor threshold (< 10ppb). DIPE in drinking water may affect the use of drinking water. To the extent that DIPE may impact the use of surface or groundwater, appropriate assessment and remediation (if necessary) should be implemented.

III. TEST PLAN RATIONALE AND DATA SUMMARY

All DIPE test data identified within this document were developed using the parent substance. The data used to characterize the alga endpoints were developed using the ECOSAR computer model (ECOSAR, 2004) provided within EPI SuiteTM (2000). This model applies an equation for neutral organics to estimate aquatic toxicity and is therefore considered appropriate to estimate aquatic toxicity for DIPE. As further justification, calculated data from this model for the fish and daphnid effect endpoints are consistent with the measured data for these organism types, supporting its use to provide adequate data for the alga endpoints.

Data used to characterize the various physicochemical, environmental fate, and environmental and mammalian health endpoints are described below.

A. Physicochemical Data

Measured DIPE physicochemical data from the literature are listed in Table 1. The log Kow value referenced as Hansch *et al.* (1995) was selected over the value referenced by Eadsforth (1983), because the primary reference cited in Hansch *et al.* (1995) has been published in external peer reviewed literature, has been carefully critiqued and included in the Syracuse Research Corp., EPISUITE® data base. That value, because of its' inclusion has been widely used and is commonly accepted. The Eadsforth (1983) reference is an internal Shell Research publication with no external review.

MELTING POINT (°C)	BOILING POINT (°C at 1013 hPa)	DENSITY (g/cm³ at 20°C)	VAPOR PRESSURE (Pa at 25°C)	WATER SOLUBILITY (mg/L at 20°C)	LOG K _{ow} (25°C)
					1.52
-86.8	68.5	0.7241	19,865	8,800	(Hansch <i>et al</i> ., 1995)
(Lide <i>et al</i> ., 1997-1998)	(Lide <i>et al</i> ., 1997-1998)	(Lide <i>et al.</i> , 1997-1998)	(Daubert and Danner, 1989)	(Gerhartz <i>et al</i> ., 1987)	2.4
					Eadsforth, 1983)

Table 1. Select Physico-Chemical Properties for DIPE.

B. Environmental Fate Data

Biodegradation

Biodegradation of an organic substance by bacteria can provide energy and carbon for microbial growth. This process results in a structural change of an organic substance and can lead to the complete degradation of that substance, producing carbon dioxide and water.

The test guideline used to assess the biodegradability of DIPE was OECD 301D, Closed Bottle Biodegradation Test. This test system design uses a sealed bottle, which is appropriate considering the test material is relatively volatile. The source of the microbial inoculum used in this study was a domestic wastewater treatment facility and the inoculum was not acclimated.

DIPE did not exhibit measurable biodegradation (0%) after 28 days under the conditions of this test design (Stone and Watkinson, 1983). While these data indicate a reduced ability for DIPE to biodegrade in compartments such as soil, water, sediment, and groundwater, there are other data in the open literature which indicate that DIPE can biodegrade in the environment. A biodegradation inhibition study was also included in the test design and it showed that the test substance did not inhibit the biodegradability of the positive control substance, sodium benzoate.

Photodegradation - Photolysis

Direct photochemical degradation occurs through the absorbance of solar radiation by a chemical substance in aqueous solution. If the absorbed energy is high enough, then the resultant excited state of the chemical may undergo a transformation. A prerequisite for direct photodegradation is the ability of one or more bonds within a chemical to absorb ultraviolet (UV)/visible light in the 290 to 750 nm range. Light wavelengths longer than 750 nm do not contain sufficient energy to break chemical bonds, and wavelengths below 290 nm are shielded from the earth by the stratospheric ozone layer (Harris, 1982a).

An approach to assessing the potential for a substance to undergo photochemical degradation is to assume that degradation will occur in proportion to the amount of light wavelengths >290 nm absorbed by constituent molecules (Zepp and Cline, 1977). The

oxygen non-bonding electrons in ethers do not give rise to absorption above 160 nm, which is why pure ether solvents can be used in spectroscopic studies. Consequently, DIPE is not subject to photolytic processes in the aqueous environment.

Photodegradation - Atmospheric Oxidation

Photodegradation can be measured (US EPA, 1999a) or estimated using an atmospheric oxidation potential (AOP) model accepted by the EPA (US EPA, 1999b). Atmospheric oxidation as a result of hydroxyl radical attack is not direct photochemical degradation, but rather indirect degradation.

DIPE has the potential to volatilize to air, based on a vapor pressure of 19,865 Pa at 25°C (Daubert and Danner, 1989), where it is subject to atmospheric oxidation. In air, DIPE can react with photosensitized oxygen in the form of hydroxyl radicals (OH-). The computer program AOPWIN (atmospheric oxidation program for Microsoft Windows) (EPI SuiteTM, 2000) calculates a chemical half-life for a 12-hour day (the 12-hour day half-life value normalizes degradation to a standard day light period during which hydroxyl radicals needed for degradation are generated), based on an OH- reaction rate constant and a defined OH- concentration.

DIPE has a calculated half-life in air of 5.3 hours or 0.4 days (12-hour day), based on a rate constant of 24.34 x 10⁻¹² cm³/molecule-sec and an OH- concentration of 1.5 x 10⁶ OH-/cm³.

Stability in Water (Hydrolysis)

Hydrolysis of an organic chemical is the transformation process in which a water molecule or hydroxide ion reacts to form a new carbon-oxygen bond. Chemicals with leaving groups that have a potential to hydrolyze include alkyl halides, amides, carbamates, carboxylic acid esters and lactones, epoxides, phosphate esters, and sulfonic acid esters (Gould, 1959). The lack of a suitable leaving group renders a compound resistant to hydrolysis. DIPE is resistant to hydrolysis because it lacks a functional group that is hydrolytically reactive and Harris (1982b) identifies ether groups as generally resistant to hydrolysis. Therefore, hydrolysis will not contribute to the removal of DIPE from the environment.

Chemical Distribution In The Environment (Fugacity Modeling)

Fugacity-based multimedia modeling provides basic information on the relative distribution of a chemical between selected environmental compartments (i.e., air, soil, water, sediment, suspended sediment, and biota). Two widely used fugacity models are the EQC (Equilibrium Criterion) Level I and Level III model (Mackay, 1996; Mackay, 1998). The Mackay Level I and Level III Models do not include a groundwater compartment.

The input data required to run a Level I model include basic physicochemical parameters; distribution is calculated as percent of chemical partitioned to 6 compartments (air, soil, water, suspended sediment, sediment, biota) within a unit world. Level I data are basic partitioning data that allow for comparisons between chemicals and indicate the compartment(s) to which a chemical may to partition, based on selected physical parameters. The Level III model uses the same physical parameters as the Level I model, but also requires half-life degradation data for the air,

soil, water, and sediment compartments, as well as emission parameters for the air. water, and soil compartments.

Results of the Mackay Level I environmental distribution model (Table 2) suggest that DIPE will partition primarily to the air, >97%. These results can be largely explained by its vapor pressure, 19,865 Pa at 25°C (Daubert and Danner, 1989). In comparison, the Level III model suggests that the majority of DIPE will partition to the water compartment, 61%, followed by the air and soil compartments with approximately equal partitioning (Table 3). These results can be explained by the model parameters. including the use of default emission rates and degradation half-lives.

Table 2. Environmental distribution as calculated by the Mackay (1998) Level I fugacity model.

ENVIRONMENTAL COMPARTMENT	DIPE DISTRIBUTION* (%)
Air	97.83
Water	2.10
Soil	0.06
Sediment	<0.01
Suspended Sediment	<0.01
Biota	<0.01

^{*}Distribution is based on the following model input parameters for DIPE:

Molecular Weight 102.18

Temperature 25° C

Log Kow 1.52 Water Solubility 8,800 g/m³ Vapor Pressure 19,865 Pa Melting Point -86.8° C

Table 3. Environmental distribution as calculated by the Mackay (1998) Level III fugacity model.

ENVIRONMENTAL COMPARTMENT	DIPE DISTRIBUTION* (%)
Air	19.40
Water	61.00
Soil	19.50
Sediment	0.10

^{*}Distribution is based on the following model input parameters for DIPE:

Emission Rate of 1,000 kg/hr into each of the air, water, and soil compartments

Molecular Weight 102.18
Temperature 25° C
Log Kow 1.52

Water Solubility 8,800 g/m³ Vapor Pressure 19,865 Pa Melting Point -86.8° C

Bioaccumulation Potential

A bioconcentration factor (BCF) of 2.95 (log BCF = 0.47) for DIPE is calculated (EPI SuiteTM, 2000) using a log K_{ow} value of 1.52 (Hansch *et al.*, 1995). A BCF of 14.06 (log BCF = 1.15) is calculated (EPI SuiteTM, 2000) when a log K_{ow} value of 2.4 (Eadsforth, 1983) is used. These data indicate that DIPE has a low BCF and is not expected to bioaccumulate.

C. Aquatic Toxicity Data

Data are available to characterize the fish toxicity of DIPE. Based on measured test concentrations, 96-hour LC₅₀ toxicity values of 91.7 to 900 mg/L have been reported (Broderius and Kahl, 1985; Geiger et al., 1985; Veith et al., 1983). Fathead minnows (Pimephales promelas) were exposed to DIPE in a 96-hour flow-through experiment resulting in an LC₅₀ of 900 mg/L (Broderius and Kahl, 1985). Geiger et al. (1985) also evaluated the toxicity of DIPE to P. promelas in flow-through exposures and reported a 96-hour LC₅₀ of 476 mg/L. These values are considerably higher than the 96-hour LC₅₀ of 91.7 mg/L measured by Veith et al. (1983) in flow through exposures with P. promelas. Dawson et al. (1977) conducted static 96-hour experiments using Lepomis macrochirus and the saltwater species Menidia beryllina. The LC₅₀ values measured in these experiments were 7000 and 6600 mg/L, respectively. The test with L. macrochirus did not yield a monotonic increasing concentration-response and the data reported by Dawson et al. are not considered reliable. Bridie et al. (1979) measured a 24-hour LC₅₀ of 380 mg/L for the goldfish Carassius auratus exposed to DIPE under static conditions. However, the report lacked sufficient detail to assess study quality and determine whether the results were based on nominal or measured values.

The acute fish data are adequate for the HPV Program and range from 91 to 900 mg/L. There is also an ECOSAR (2004) calculated fish 96-hour LC_{50} of 215 mg/L, which is within the reported range of acute toxicity. This model is appropriate to estimate aquatic toxicity for this class of chemicals.

Stephenson (1983) reported a 48-hour invertebrate (*Daphnia magna*) EC_{50} toxicity value of 190.0 mg/L based on nominal test concentrations. A 48-hour daphnid LC_{50} toxicity value of 220 mg/L was estimated using the ECOSAR (2004) model.

There are no reliable measured algal toxicity data. Stephenson (1983) conducted an algal study and reported a 96h EC $_{50}$ of >1000 mg/L. Test concentrations were not measured and there is no indication in the report whether the test vessels were sealed. The reported LC $_{50}$ value may reflect a loss of test substance by volatilization if the flasks were not tightly sealed. A green alga 96-hour EC $_{50}$ toxicity value of 135 mg/L and a 96-hour chronic value of 10.2 mg/L were calculated (ECOSAR, 2004).

Conclusion

Based on the available data, DIPE presents a low order of acute toxicity to aquatic species. Although the reported invertebrate study was assessed as reliable, the results are based on nominal values. Therefore, an additional study will be conducted that is comparable in quality to the fish studies based on measured results. Although a study was reported for an alga species, it was assessed as unreliable. Therefore, an alga study will be conducted that reports measured results.

Table 4. Aquatic Toxicity Values for DIPE.

ENDPOINT	MEASURED VALUE(S) (mg/L)	CALCULATED VALUE (mg/L)
Fish 96-hr LC ₅₀	91.7 (Veith, 1983) to 900 (Broderius and Kahl, 1985)	214.1 (ECOSAR, 2004)
Daphnid 48-hr EC ₅₀ /LC ₅₀	190.0* (Stephenson, 1983)	221.9 (ECOSAR, 2004)
Alga 96-hr EC ₅₀	No reliable measured data	134.9 (ECOSAR, 2004)
Alga 96-hr ChV	No reliable measured data	10.2 (ECOSAR, 2004)

^{*} nominal value

D. Human Health Effects Data

Mammalian toxicity data for DIPE are summarized below in Table 5. Each endpoint is discussed in the following sections. Additional data for studies beyond those required in the HPV Program are also presented below.

Acute Toxicity

DIPE is low in acute toxicity to mammals. Single lethal dosages/concentrations of DIPE to laboratory animals range from 4.5 to 16.5 g/kg (LD_{50}) for oral exposure, greater than 20 mL/kg (LD_{50}) for dermal exposure and between 8000 and 16000 ppm (33.6 and 67.2 mg/L) for inhalation exposure (Kimura *et al.*, 1971). Typical clinical signs in acutely poisoned animals are the result of central nervous system (CNS) effects and include spasmodic movement, tremors, convulsions, incoordination and unsteadiness, narcosis, anesthesia, coma and respiratory depression (Machle *et al.*, 1939). At necropsy,

lesions in affected animals following acute exposure include gastrointestinal irritation, visceral congestion and pulmonary edema (Machle *et al.*, 1939).

There are no reports of human systemic toxicity associated with acute DIPE exposure.

Conclusion

DIPE is of a low order of acute oral, dermal, and inhalation toxicity with LD₅₀ values in excess of 2000 mg/kg and an inhalation LC₅₀ >20 mg/kg.

Genotoxicity

In vitro

DIPE is not genotoxic in a number of *in vitro* assays. DIPE did not induce reverse gene mutation in bacterial tester strains S. *typhimurium* (6 strains), and *E. coli* (3 strains) or mitotic gene conversion in the yeast *S. cervisiae* JD1, with or without metabolic activation. DIPE did not induce chromosome damage in cultured rat liver (RL₄) or CHO cells (Brooks *et al*, 1988).

Conclusion

DIPE is not genotoxic in vitro and DIPE is not considered a mutagenic hazard.

Repeated Dose Toxicity

A number of subchronic inhalation studies exist for DIPE. Rats were exposed to 0, 480, 3300, and 7100 ppm DIPE for 6 hours/day, 5 days/wk for 13 weeks. Increases in liver and kidney weights were seen at 3300 and 7100 ppm in both males and females. Some evidence of increased incidence of hyaline droplets in kidney proximal tubules was observed in high dose males only. No effects on serum chemistry, hematology, or pathology were noted at any dose level. The no observed adverse effect level (NOAEL) for this study was 480 ppm (Dalbey and Feuston, 1996).

Guinea pigs, rabbits and monkeys were subchronically exposed to DIPE under the following conditions: 0.1% (1000 ppm) for 3 hours/day, 0.3% (3000 ppm) for 2 hours/day, or 1.0% (10000 ppm) for 1 hour/day for 20 exposures. In addition, rabbits and monkeys received 3.0% (30000 ppm) for 10 exposures. No major deleterious effects were noted in any animals. Some CNS depression was seen in monkeys at 1% and guinea pigs at 3.0% DIPE. Anesthesia with a prompt recovery upon cessation of exposure and some body weight loss were seen at 3% in monkeys (Machle *et al.*, 1939).

Reproductive and Developmental Toxicity

Rats were administered 0, 430, 3095, and 6745 ppm DIPE for 6 hours/day on gestations days 6-15. Maternal effects at the high dose included increased salivation and lacrimation during and immediately following exposure. A slight decrease in food consumption was noted at 3095 and 6745 ppm. A concentration-related increase in the incidence of rudimentary ribs was observed (statistically significant at 3095 and 6745 ppm), but the significance of this finding is not known. The NOAEL for both maternal and developmental effects under conditions of this study was 430 ppm (Dalbey and Feuston, 1996).

No changes in reproductive organ weights and structure or sperm and spermatid number at any dose group were noted in rats exposed to 0, 480, 3300, and 7100 ppm DIPE for 6 hours/day, 5 days/wk for 13 weeks (Dalbey and Feuston, 1996).

Conclusion

NOAEL for both maternal and developmental effects in the developmental study was 430 ppm and the NOAEL for the subchronic study was 480 ppm.

Additional Human Health Effects Data

Irritation

DIPE is mildly irritating to skin, eye and respiratory mucosa in limited animal studies. Single or repeated exposures of DIPE to rabbit skin produced some reddening and dermatitis that subsided after 2 weeks. Direct application to rabbit eyes produces trace effects (Machle *et al.*, 1939).

Irritation effects associated with DIPE vapor have been reported in human volunteers. Silverman *et al.* (1946) reported that 35% of humans exposed to DIPE vapor at a concentration of 300 ppm objected to the unpleasant odor of the solvent. At 800 ppm for 5 minutes, most subjects reported irritation of the eyes and nose, and the most sensitive reported respiratory discomfort. Concentrations above 1000 ppm DIPE resulted in complaints of strong irritation to the eyes and respiratory tract. Another study evaluated 5-minute exposures in volunteers and found only slight irritation of the nose at 400 ppm progressing to slight irritation of nose, eyes and respiratory tract at 800 ppm (Hine *et al.*, 1955). Subjective reporting and the potential influence of odor complicate interpretation of these studies.

Conclusion

DIPE is not a skin irritant but prolonged/repeated contact may cause defatting of the skin, which can lead to dermatitis. The vapours and liquid may be irritating to the eyes at 800 ppm, but not at 300 ppm.

Sensitization

No *in vivo* skin sensitization studies are available for DIPE. A mathematical model for predicting sensitization based on chemical reactivity suggests a lack of sensitization potential for DIPE (Wass and Belin, 1990).

Conclusion

DIPE is not expected to be a skin sensitizer.

Neurotoxicity

Subchronic neurotoxicity potential was evaluated in rats exposed to 0, 450, 3250, or 7060 ppm DIPE by inhalation for 5 days/wk for 13 weeks. Functional observational battery (FOB) and motor activity was determined at 0, 4, 8, and 13 weeks. Minor decreases in FOB activity at 450 and 7060 ppm and in figure eight motor activity at 7060 ppm were noted at week 4. Some increases in figure eight motor activity were observed in the 450 ppm group at week 4. These changes were not seen at later time points and other FOB parameters and clinical signs were unaffected by treatment. No alterations were noted in microscopic examination of brain, spinal cord, dorsal root ganglia, and sciatic nerve (Rodriguez and Dalbey, 1997).

Conclusion

Inhalation exposures to DIPE at concentrations as high as 7060 ppm for 13 weeks resulted in few observable effects on the nervous system.

Carcinogenicity

DIPE was assessed in a long term gavage study in rats at dose levels of 0, 250 and 1000 mg/kg/day, 4 days a week for 78 weeks (Belpoggi et al, 2002; no robust summary provided). There were a number of deficiencies in design and reporting of this non-conventional study, which made interpretation difficult. These included the maintenance of the animals to their natural death (usually studies are terminated at the end of a predetermined exposure period and/or survival level), lack of detail in reporting statistical significance of tumor incidence (e.g., combined lymphoma and leukemia), and limited reporting of survival. There was no indication of or comparison to historical control data. For these reasons the findings and the significance of such are questionable and are considered equivocal.

All animals were observed until spontaneous death and the experiment ended after 163 weeks. There were no significant differences between treated and control groups in daily food or water consumption, body weight, behavior or non-neoplastic pathological changes. Survival was decreased in treated males compared to controls between the 56th and 88th weeks of age. A statistically significant increase in total malignant tumours was reported in males at the low dose only and a significant trend in females of both treated groups. The incidence of carcinomas of the ear duct in males was statistically significant but was not dose related. A statistically significant increase in combined hemo/lympho-reticular neoplasias (% tumor-bearing animals) was seen in males and females at both dose levels. The increase in tumor incidence was not statistically significant for low dose males but the trend was. Significance was not reported for individual tumours (% tumor bearing animals) or for individual types of lymphoma or leukemia.

Conclusion

The findings of this "lifetime" carcinogenicity study in rats are equivocal.

 Table 5.
 Mammalian Toxicity Endpoint Summary for DIPE.

TOXICIT	Y ENDPOINT	RESULTS	REFERENCE
	Inhalation	Low toxicity	Kimura <i>et al.</i> , 1971; Machle <i>et al.,</i> 1939
Acute	Oral	Low toxicity	Kimura <i>et al.</i> , 1971; Machle <i>et al.</i> , 1939
	Dermal	Low toxicity	Kimura et al., 1971; Machle et al., 1939
	Skin	Minimal irritant	Machle et al., 1939
Irritation	Eye	Minimal irritant	Machle et al., 1939
	Respiratory	Sensory irritant	Hine et al., 1955; Silverman et al., 1946
Sensitization		Negative in vitro sensitizer	Wass and Belin, 1990
Repeated Dose		Liver and kidney effects	Dalbey and Feuston, 1996
Reproductive		No effects on reproductive organ structure or sperm/spermatid number.	Dalbey and Feuston, 1996
Developmental		Equivocal developmental effect at maternal effect level	Dalbey and Feuston, 1996
Neurotoxicity		Minor reversible effect on CNS	Rodriguez and Dalbey, 1997
Genotoxicity In vitro mutation In vitro chromosome aberration		Negative	Brooks <i>et al.</i> , 1988
		Negative	Brooks <i>et al.</i> , 1988

V. TEST PLAN SUMMARY

A search for existing studies/information identified data to characterize all endpoints under the U.S. EPA HPV Program for DIPE (Table 6). However, the acute invertebrate and alga toxicity endpoints did not have measured data that were equivalent in quality to the data used to characterize the fish acute toxicity endpoint. Therefore, studies will be conducted to develop new data for these endpoints. A dossier containing the robust summaries of the data presented in this test plan is provided in the Appendix.

Table 6. DIPE Data Availability and Adequacy for Endpoints in the HPV Program.

Endpoint	Data Availability	Reference	First Author/ Source	Date
Physical/Chemical Propertie	IS			
2.1 Melting Point	Adequate measured	22		
2.2 Boiling Point	Adequate measured	22	Lide et al.	1997
2.3 Density	Adequate measured	22		
2.4 Vapor Pressure	Adequate measured	6	Daubert and Danner	1989
2.5 Partition Coefficient	Adequate	15	Hansch et al.	1995
	measured	8	Eadsforth	1983
2.6 Water Solubility	Adequate measured	12	Gerhartz et al.	1987
Environmental Fate				
	Technical	16	Harris	1982a
3.1 Photodegradation	discussion (direct)	35	Zepp and Cline	1977
(direct and indirect)	Computer model (indirect)	9	EPI Suite [™]	2000
3.2 Stability in Water	Technical	13	Gould	1959
	discussion	17	Harris	1982b
3.3 Transport between Environmental	Computer	24	Mackay	1998
Compartments (Level I and III)	model	25	Mackay et al.	1996
3.4 Biodegradation	Adequate measured	30	Stone and Watkinson	1983
3.5 Bioaccumulation	Computer model	9	EPI Suite™	2000

Endpoint	Data Availability	Reference	First Author/ Source	Date
Environmental Toxicity				
	Adequate measured	3	Broderius and Kahl	1985
	Adequate measured	11	Geiger et al.	1986
4.1 Acute/prolonged toxicity to	Adequate measured	33	Veith <i>et al.</i>	1983
Fish	Computer model	10	ECOSAR	2004
	Measured (not reliable)	2	Bridie et al.	1979
	Measured (not reliable)	7	Dawson	1975/77
4.2 Acute Toxicity to Aquatic	Measured (nominal)	29	Stephenson	1983
Invertebrates	Computer model	10	ECOSAR	2004
	Measured (not reliable)	29	Stephenson	1983
4.3 Toxicity to Aquatic Plants	Computer model	10	ECOSAR	2004

	ndpoint	Data Availability	Reference	First Author/ Source	Date
Mammalian '	Foxicity				
5.1.1. Acute (Oral Taxiaity	Adequate	20	Kimura <i>et al.</i>	1971
5.1.1. Acute (Dial Toxicity	measured	23		
5.1.2 Acute Ir	halation Toxicity	Adequate measured	23	Machle <i>et al</i> .	1939
5.1.3 Acute D	ermal Toxicity	Adequate measured	23		
5.2 Sensitizat	ion	Adequate measured	34	Wass and Belin	1990
		Adequate	5	Dalbey and Feuston	1996
5.3 Repeated	Dose Toxicity	measured	27	Rodriguez and Dalbey	1997
5.4 Genetic Toxicity	Bacterial mutagenicity	Adequate	4	Brooks et al.	1998
In Vitro Chromosomal damage		measured	4	BIOOKS & al.	1990
5.5 Developm Toxicity/T	nental eratogenicity	Adequate measured	5	Dalbey and Feuston	1996
5.6 <u>Additional Information</u> – Sensory Irritation in Humans		Adaquata	18	Hine et al.	1955
		Adequate measured	28	Silverman <i>et</i> al.	1946

V. REFERENCES

- 1. Belpoggi F, Soffritti M, Minardi F, Bua L, Cattin E and Maltoni C (2002). Results of long-term carcinogenicity bioassays on tert-amyl-methyl-ether (TAME) and Diisopropyl ether (DIPE) in rats. Ann. N.Y. Acad. Sci. 982, 70-86.
- 2. Bridie A, Wolff C and Winter M (1979). The acute toxicity of some petrochemicals to goldfish. Water Res. 13, 623-626.
- 3. Broderius S and Kahl M (1985). Acute toxicity of organic chemical mixtures to the fathead minnow. Aguat. Toxicol. 6, 307-322.
- 4. Brooks T, Meyer A and Hutson D (1988). The genetic toxicology of some hydrocarbon and oxygenated solvents. Mutagenesis 3(3), 227-232.
- 5. Dalbey W and Feuston M (1996). Subchronic and developmental toxicity studies of vaporized disopropyl ether. J. Toxicol. Environ. Health 49, 29-43.
- 6. Daubert T and Danner R (1989). Physical and thermodynamic properties of pure chemicals data compilation. Taylor and Francis, Washington, D.C., USA.
- 7. Dawson G, Jennings A, Drozdowski D and Rider E (1975/77). The acute toxicity of 47 industrial chemicals to fresh and saltwater fishes. J. Haz. Mat. 1, 303-318.
- 8. Eadsforth C (1983). Isopropyl ether: determination of the n-octanol/water partition coefficient using a reverse-phase HPLC method. Report # SBGR.83.131. Shell Research Limited, Sittingbourne Research Centre, Sittingbourne, Kent, England.
- 9. EPI SuiteTM (2000). Estimation Program Interface Suite, v3.12. Syracuse Research Corporation, Syracuse, NY, USA.
- 10. ECOSAR v0.99h (2004) in EPI Suite[™] (2000). Estimation Program Interface Suite, v3.12. Syracuse Research Corporation, Syracuse, NY, USA.
- Geiger D, Poirier S, Brooke L and Call D (eds.) (1986). Acute Toxicities of Organic Chemicals to Fathead Minnows (*Pimephales promelas*), Vol. 3. Center for Lake Superior Environmental Studies, Univ. of Wisconsin- Superior, Superior, WI, USA.
- Gerhartz W, Yamamoto Y, Kaudy L, Rounsaville J and Schulz G (eds.) (1987).
 Ullmann's Encyclopedia of Industrial Chemistry. Vol. A 10. 5th Edition. VCH Publishers, New York, NY, USA.
- 13. Gould E (1959). Mechanism and Structure in Organic Chemistry. Holt, Reinhart and Winston, New York, NY, USA.
- GTPZAB (Gigiena Truda Professional nye Zabolevania) (1975). Vol. 19(10), 55.
 (Cited in Lewis R.J. Sr. (1992). Sax's Dangerous Properties of Industrial Materials. Eighth Edition. Van Nostrand Reinhold, New York, NY, USA).

- 15. Hansch C, Leo A and Hoekman D (1995). Exploring QSAR Hydrophobic, Electronic and Steric Constants. p. 6. ACS Professional Reference Book, American Chemical Society, Washington, DC, USA.
- 16. Harris J (1982a). Rate of aqueous photolysis. In: Handbook of Chemical Property Estimation Methods, Lyman W, Reehl W and Rosenblatt D (eds.), Chapter 8. McGraw-Hill Book Company, New York, NY, USA.
- 17. Harris J (1982b). Rate of hydrolysis. In: Handbook of Chemical Property Estimation Methods, Lyman W, Reehl W and Rosenblatt D (eds.), Chapter 7. McGraw-Hill Book Company, New York, NY, USA.
- 18. Hine C, Anderson H and Kodama J (1955). Sensory thresholds of certain Shell organic solvents, Progress Report 1, Report to Shell Development Company, November 15, 1955, UC Report #247.
- Jackson D (1933). The pharmacological action of isopropyl ether. Soc. Proc. of Amer. Soc. Pharm. Exper. Therap., 24th Annual Meeting. J. Pharm. Exper. Therap. 48, 267.
- 20. Kimura E, Ebert D and Doge P (1971). Acute toxicity and limits of solvent residues for sixteen organic solvents. Toxicol. Appl. Pharmacol. 19, 699-704.
- 21. Klimisch H, Andreae M and Tillmann U (1997). A systematic approach for evaluating the quality of experimental toxicological and ecotoxicological data. Regulatory Toxicol. Pharmacol. 25, 1-5.
- 22. Lide D, *et al.* (eds.) (1997-1998). CRC Handbook of Chemistry and Physics. 78th Edition. CRC Press, New York, NY, USA.
- 23. Machle W, Scott E and Treon J (1939). The physiological response to isopropyl ether and to a mixture of isopropyl ether and gasoline. J. Ind. Hyg. Toxicol. 21, 72-96.
- 24. Mackay D (1998). Fugacity-Based Environmental Equilibrium Partitioning Model, Version 2.1 (16-bit). Environmental Modelling Centre, Trent University, Ontario, Canada.
- 25. Mackay D, Di Guardo A, Paterson S and Cowan C (1996). Evaluating the environmental fate of a variety of types of chemicals using the EQC model. Environ. Toxicol. Chem. 15, 1627-1637.
- 26. Pavlova L, Lagunova V and Imanov R (1975). Data for the determination of the maximum allowable concentration of diisopropyl ether in the air of a workplace environment. Tr. Azerb. Nauchno-Issled. Inst. Gig. Tr. Prof. Zabol. 10, 55-57. [translation from Russian]
- 27. Rodriguez S and Dalbey W (1997). Subchronic neurotoxicity of vaporized diisopropyl ether in rats. International J. Toxicol. 16, 599-610.

- 28. Silverman L, Schulte F and First M (1946). Further studies on sensory response to certain industrial solvent vapors. J. Ind. Hyg. Toxicol. 28(6), 262-266.
- 29. Stephenson R (1983). Isopropyl ether: Acute toxicity to *Daphnia magna* and *Selenastrum capricornutum*. Report # SBGR.83.215, Shell Research Ltd. Sittingbourne Research Centre, Sittingbourne, Kent, England.
- 30. Stone C and Watkinson R (1983). Isopropyl ether: An assessment of ready biodegradability. Report # SBGR.83.428. Shell Biosciences Laboratory, Sittingbourne Research Centre, Sittingbourne, Kent, England.
- 31. U.S. EPA (1999a). Determining the Adequacy of Existing Data. OPPT, EPA, Washington, DC, USA.
- 32. U.S. EPA (1999b). The Use of Structure-Activity Relationships (SAR) in the High Production Volume Chemicals Challenge Program. OPPT, EPA, Washington, DC, USA.
- 33. Veith G, Call D and Brooke L (1983). Structure-toxicity relationships for the fathead minnow, *Pimephales promelas*: Narcotic industrial chemicals. Can. J. Fish. Aquat. Sci. 40, 743-748.
- 34. Wass U and Belin L (1990). An *in vitro* method for predicting sensitizing properties of inhaled chemicals. Scand. J. Work Environ. Health 116, 208-214.
- 35. Zepp R and Cline D (1977). Rate of direct photolysis in the aqueous environment. Environ. Sci. Technol. 11, 359-366.

APPENDIX

Data Dossier (Robust Study Summaries) for Diisopropyl Ether (CAS No. 108-20-3)

(The format of the following dossier is taken largely from IUCLID, International Uniform Chemical Information Database)

1. General Information

ld 108-20-3 Date 12.16.2005

Data Dossier (Robust Study Summaries) for Diisopropyl Ether (CAS No. 108-20-3)

1.1 GENERAL SUBSTANCE INFORMATION

Purity type

Substance type

organic

Physical status

liquid

Purity Color Odor

:

1.2 SYNONYMS AND TRADENAMES

2,2'-Oxybis-propane

2,2'-Oxybispropane

Propane, 2,2'-oxybis-

2-Isopropoxy propane

2-Isopropoxypropan

2-Isopropoxypropane

Dipropyloxid

IPE

DIPE

Isopropyl ether

Isopropylether

Diisopropyl ether

Diisopropylether

2. Physico-Chemical Data

ld 108-20-3 Date 12.16.2005

2.1 **MELTING POINT**

Value

= -86.8 °C

Method

: other: not specified

GLP

: no data

Test substance

: Diisopropyl ether (CAS # 108-20-3)

Test substance

: CAS No. 108-20-3; diisopropylether; purity is unknown.

Reliability

: (2) valid with restrictions

The CRC Handbook of Chemistry and Physics is a peer reviewed publication. This robust summary has a reliability rating of 2 because there

is insufficient information available on the method and analytical procedure.

Flag

Critical study for SIDS endpoint

Reference Number

(14)

BOILING POINT 2.2

Value

= 68.5 °C at 1013 hPa

Method

other: not specified

GLP

no data

Test substance

: Diisopropyl ether (CAS # 108-20-3)

Test substance

CAS No. 108-20-3; diisopropylether; purity is unknown.

Reliability

(2) valid with restrictions

The CRC Handbook of Chemistry and Physics is a peer reviewed

publication. This robust summary has a reliability rating of 2 because there is insufficient information available on the method and analytical procedure.

Flag

: Critical study for SIDS endpoint

Reference Number

(14)

DENSITY 2.3

Type

density

Value

: = .7241 g/cm³ at 20 °C

Method

: other: not specified

GLP

: no data

Test substance

: Diisopropyl ether (CAS # 108-20-3)

Test substance

: CAS No. 108-20-3; diisopropylether; purity is unknown.

Reliability

(2) valid with restrictions

The CRC Handbook of Chemistry and Physics is a peer reviewed

publication. This robust summary has a reliability rating of 2 because there is insufficient information available on the method and analytical procedure.

Flag

Critical study for SIDS endpoint

Reference Number

(14)

2.4 **VAPOUR PRESSURE**

Value

: = 198.65 hPa at 25 °C

GLP

: no data

Test substance

: Diisopropyl ether (CAS # 108-20-3)

Method

: Method not specified.

Test substance

: CAS No. 108-20-3; diisopropylether; purity is unknown.

Reliability

(2) valid with restrictions

This robust summary has a reliability rating of 2 because the data were not

reviewed for quality, however, the reference is from a peer-reviewed

2. Physico-Chemical Data

ld 108-20-3 Date 12.16.2005

handbook.

Flag

Critical study for SIDS endpoint

Reference Number

(4)

2.5 **PARTITION COEFFICIENT**

Partition coefficient

Log pow

octanol-water = 1.52 at 25 °C

other (measured)

Method **GLP**

: no data

Test substance

: Diisopropyl ether (CAS # 108-20-3)

Method

: Method not specified. : CAS No. 108-20-3; diisopropylether purity is unknown.

Test substance

: (2) valid with restrictions

Reliability

The value cited by the authors is a measured and preferred value. This robust summary has a reliability rating of 2 because there is insufficient

information available on the method and analytical procedure.

Flag

Critical study for SIDS endpoint

Reference Number

(10)

Partition coefficient

Log pow

octanol-water = 2.4 at °C other

Method **GLP**

Nο

Test substance

Diisopropyl ether (CAS No. 108-20-3) Indirect method by reverse-phase HPLC Log Pow = 2.4 (Pow = 250) at pH 6.7

Method Result

Test condition

The HPLC system was a reverse-phase C18-coated silica gel column, 250 mm x 5 mm id, with a mobile phase of 3 volumes methanol and 1 volume water (final pH 6.7) at a flow rate of 1 mL/min. 100 µL samples of an approximate 1 mg/mL solution in the mobile phase were injected, and the emergence of the material was observed using refraction index detection.

Thirty-one reference compounds were used to generate the linear relationship between log k (k = capacity factor) and log Pow. Using the HPLC retention time for the peak of the test substance, the log k was determined, and the log Pow value was calculated using the linear equation developed from the reference compounds.

Log Pow was determined according to the following calculations:

Retention time (RT), min = 5.7

Capacity factor, k = 0.87, k = (RT_{cmpd} - RT_{unretained std})/RT_{unretained std}

log k = -0.06

linear equation: $\log k = -0.930 + 0.357 \log Pow$

Reliability **Reference Number** (1) Valid without restrictions

(4.2)

2.6 SOLUBILITY IN DIFFERENT MEDIA

Solubility in

Water

Value

= 8800 mg/l at 20 °C

GLP

no data

Test substance

Diisopropyl ether (CAS # 108-20-3)

Test substance

Reliability

CAS No. 108-20-3; diisopropylether; purity is unknown.

(2) valid with restrictions

The Ullmann's Encyclopedia of Industrial Chemistry is a peer reviewed publication. This robust summary has a reliability rating of 2 because there is insufficient information available on the method and analytical procedure.

Flag

Critical study for SIDS endpoint

Reference Number

(8)

(6)

3.1 PHOTODEGRADATION

Type

Air

Conc. of substance

at 25 °C

INDIRECT PHOTOLYSIS

Sensitizer

: OH

Conc. of sensitizer

: 1.5E6 OH- radicals/cm3

Rate constant

 $= .00000000002434 \text{ cm}^3/(\text{molecule*sec})$

Degradation

: = 50 % after 5.3 hour(s)

Method

: other (calculated): Calculated values using AOPWIN version 1.89, a

subroutine of the computer program EPIWIN version 3.12

Test substance

Diisoproply Ether (CAS # 108-20-3)

Method

Calculated values using AOPWIN version 1.89, a subroutine of the

computer program EPIWIN version 3.12

Indirect photodegradation, or atmospheric oxidation potential, is based on the structure-activity relationship methods developed by R. Atkinson under the following conditions:

Temperature: 25°C Sensitizer: OH- radical

Concentration of Sensitizer: 1.5E6 OH- radicals/cm3

Remark

DIPE has the potential to volatilize to air, based on a vapor pressure of 19,865 Pa at 25°C (Daubert and Danner, 1989), where it is subject to atmospheric oxidation. In air, DIPE can react with photosensitized oxygen in the form of hydroxyl radicals (OH-). The computer program AOPWIN (atmospheric oxidation program for Microsoft Windows) (EPIWIN, 2000) calculates a chemical half-life for a 12-hour day (the 12-hour day half-life value normalizes degradation to a standard day light period during which hydroxyl radicals needed for degradation are generated), based on an OHreaction rate constant and a defined OH- concentration.

DIPE has a calculated half-life in air of 5.3 hours or 0.4 days (12-hour day), based on a rate constant of 24.34 E-12 cm3/molecule*sec and an OH-

concentration of 1.5 E5 OH-/cm3.

Reliability

(2) valid with restrictions

The value was calculated based on chemical structure as modeled by EPIWIN. This robust summary has a reliability rating of 2 because the data

are calculated and not measured.

Flag

Reference Number

Critical study for SIDS endpoint

Test substance

: Diisoproply Ether (CAS # 108-20-3)

Method Remark Technical discussion

: Direct photochemical degradation occurs through the absorbance of solar radiation by a chemical substance in aqueous solution. If the absorbed energy is high enough, then the resultant excited state of the chemical may undergo a transformation. A prerequisite for direct photodegradation is the ability of one or more bonds within a chemical to absorb ultraviolet (UV)/visible light in the 290 to 750 nm range. Light wavelengths longer than 750 nm do not contain sufficient energy to break chemical bonds, and wavelengths below 290 nm are shielded from the earth by the stratospheric

ozone layer (Harris, 1982a).

An approach to assessing the potential for a substance to undergo photochemical degradation is to assume that degradation will occur in proportion to the amount of light wavelengths >290 nm absorbed by constituent molecules (Zepp and Cline, 1977). The oxygen non-bonding electrons in ethers do not give rise to absorption above 160 nm, which is

why pure ether solvents can be used in spectroscopic studies.

Consequently, DIPE is not subject to photolytic processes in the aqueous

environment.

ld 108-20-3 Date 12.16.2005

Reliability

: (2) valid with restrictions

This robust summary has a reliability of 2 because it is a technical

discussion and not a study.

Flag

Reference Number

: Critical study for SIDS endpoint

(27)

3.2 STABILITY IN WATER

Type GLP : abiotic : no data

Test substance

: Diisoproply Ether (CAS # 108-20-3)

Method

: Technical discussion

Result

: Hydrolysis of an organic chemical is the transformation process in which a water molecule or hydroxide ion reacts to form a new carbon-oxygen bond. Chemicals with leaving groups that have a potential to hydrolyze include alkyl halides, amides, carbamates, carboxylic acid esters and lactones, epoxides, phosphate esters, and sulfonic acid esters (Gould, 1959). The lack of a suitable leaving group renders a compound resistant to hydrolysis. DIPE is resistant to hydrolysis because it lacks a functional group that is hydrolytically reactive and Harris (1982b) identifies ether groups as

generally resistant to hydrolysis. Therefore, hydrolysis will not contribute to

the removal of diisoproply ether from the environment.

Reliability

(2) valid with restrictions

This robust summary has a reliability of 2 because it is a technical

discussion and not a study.

Flag

Reference Number

Critical study for SIDS endpoint

(9) (11)

3.3 TRANSPORT BETWEEN ENVIRONMENTAL COMPARTMENTS

Type

: fugacity model level I

Media Method other: air - biota - sediment(s) - soil - waterCalculation according Mackay, Level I

Remark

: Physicochemical data used in the calculation:

Parameter

Value w/ Units

Molecular Weight = 102.18 Temperature = 25° C

Log Kow = 1.52

Water Solubility = 8,800 g/m3 Vapor Pressure = 19,865 Pa Melting Point = -86.8° C

Result

: Using the Mackay Level I calculation, the following distribution is predicted for disoproply ether:

%Distribution Compartment

97.83 Air 2.10 Water

0.06 Soil <0.01 Sediment

<0.01 Suspended Sediment

<0.01 Biota

Test substance Reliability Diisoproply Ether (CAS # 108-20-3)

(2) valid with restrictions

This robust summary has a reliability rating of 2 because the data are

calculated.

ld 108-20-3 Date 12.16.2005

Flag

Reference Number

: Critical study for SIDS endpoint

(17)

Type Media : fugacity model level III

: air - sediment(s) - soil - water

Method : Level III simulation using the Mackay Multimedia Environmental Model

(Mackay, 2001)

Test substance Method : Diisopropyl Ether (CAS No. 108-20-3)

: Level III simulation using the Mackay Multimedia Environmental Model (Mackay, 2001). Mass balances are calculated for the four bulk media of air (gas + aerosol), water (solution + suspended sediment + biota), soil, (solids + air + water), and sediment (solids + pore water). Equilibrium exists within, but not between media. Physical-chemical properties are used to quantify a chemical's behavior in an evaluative environment. Three types of chemicals are treated in this model: chemicals that partition into all media (Type 1), non volatile chemicals (Type 2), and chemicals with zero, or nearzero, solubility (Type 3). The model can not treat ionizing or speciating substances. The Level III model assumes a simple, evaluative environment with user-defined volumes and densities for the following homogeneous environmental media (or compartments): air, water, soil, sediment, suspended sediment, fish and aerosols.

This model provides a description of a chemical's fate including the important degradation and advection losses and the intermedia transport processes. The distribution of the chemical between media depends on how the chemical enters the system, e.g. to air, to water, or to both. This mode of entry also affects persistence or residence time.

The rates of intermedia transport are controlled by a series of 12 transport velocities. Reaction half-lives are requested for all 7 media. The advective residence time selected for air also applies to aerosols and the residence time for water applies to suspended sediment and fish. The advective residence time of aerosols, suspended sediment and fish cannot be specified independently of the air and water residence times.

Result

: Output

	Mass%	Half life(hr)	Emissions(kg/hr)
Air	19.4	25.2	1000
Water	61.0	360	1000
Soil	19.5	720	1000
Sediment	0.1	3240	0

Test condition

: Physchem Inputs

Molar Mass = 102.18 Data Temperature = 25 °C Water Solubility = 8800 mg/l exp. Vapour Pressure = 19865 Pa exp.

Log Kow = 1.52 exp.

Melting Point = -86.8 °C exp.

Reaction Half Lives in hours (if not available they can be predicted using

EPIWIN)
Air (gaseous)
Water (no sus

25.2

Water (no susp. part.) 360
Bulk Soil 720
Bulk Sediment 3240
Suspended Particles 360

Fish 360 Aerosol 25.2

Environmental Properties (EQC standard environment)

Dimensions (all defaults)
Densities (all defaults)

ld 108-20-3 Date 12.16.2005

Organic carbon & Advection (all defaults)

Transport Velocities (all defaults)

Emission and Inflows (defaults used)

Air 1000 kg/hr Water 1000 kg/hr Soil 1000 kg/hr Sediment 0 kg/hr

Conclusion

: The majority of DIPE is calculated to partition into the water phase, with smaller but significant amounts into air and soil, based on the modeling parameters used in this calculation. DIPE is considered to be a Type 1 chemical with potential to partition into all environmental compartments.

Reliability

(2) valid with restrictions

This robust summary has a reliability rating of 2 because the data are

calculated.

Flag

Reference Number

Critical study for SIDS endpoint

(16) (18) (19) (20)

3.4 **BIODEGRADATION**

Type

Inoculum

Contact time

Degradation

Result

Method Year

GLP Test substance

Resuit

: aerobic

activated sludge, domestic

28 day(s)

: 0% after 28 days

: not readily biodegradable

OECD Guide-line 301 D "Ready Biodegradability: Closed Bottle Test"

1982

Diisopropyl Ether (CAS No. 108-20-3)

Test substance was not readily biodegradable. After 28 days, the test substance exhibited no measurable biodegradation. By day 5, >60% biodegradation of positive control was observed, which meets the guideline requirement. No excursions from the testing guideline were noted. The inhibition study showed that the test substance did not inhibit the biodegradability of the positive control substance, sodium benzoate.

Sample	% Degradation* (day 28)	Mean % Degradation (day 28)
Test Substance	0.0, 0.0	0.0
Na Benzoate * duplicate data	65.0, 73.0	69.0

Mean oxygen concentrations (mg/L) of duplicate test systems:

Day 0

Mineral Salts Control = 8.85

Blank = 8.8

Na Benzoate = 8.95

Test Substance = 8.9 (single test system)

Test Substance + Na Benzoate = 8.9* (single test system)

Day 5

Mineral Salts Control = 9.0

Blank = 8.8

Na Benzoate = 5.7

Test Substance = 8..85

Test Substance + Na Benzoate = 5.8

ld 108-20-3 Date 12.16.2005

Day 15

Mineral Salts Control = 8.75

Blank = 8.65Na Benzoate = 4.9 Test Substance = 8.55

Test Substance + Na Benzoate = 4.9

Day 28

Mineral Salts Control = 8.65

Blank = 7.05Na Benzoate = 3.6 Test Substance = 8.3

Test Substance + Na Benzoate = 4.15

Test condition

The inoculum source was the Sittingbourne Sewage works in Kent. England, and was prepared according to methods described in the OECD 301D guideline. The test substance was added to the test medium by direct addition at a concentration of 3.0 mg/L. Test systems were incubated at 20 ± 1 °C and biodegradation was determined by measuring the oxygen concentration on days 5, 15, and 28. Each sampling of the test substance and control was conducted in duplicate. The theoretical oxygen demand was 2.82 mg O2 per mg test substance and a theoretical carbon dioxide (CO2) evolution of 2.59 mg CO2 per mg test substance. Sodium benzoate was used as the positive control.

The purity of the test substance was not supplied, but the infra-red spectrum of the test substance matched a published standard (density = 0.723 to 0.726 kg/L). The test substance was stored in the dark at ambient temperature. Nitrogen was blown over the surface of the material when the container was opened and exposed to air in order to minimize peroxide formation.

Conclusion

Diisopropyl ether is not readily biodegradable and it did not significantly inhibit the biodegradability of the test substance in an inhibition test.

Reliability

Reference Number

: (1) valid without restriction

(22)

3.5 **BIOACCUMULATION**

Species

other: see remark

Exposure period **BCF**

at 25 °C

= 2.95

Method

other: calculation

GLP

Test substance Remark

Diisopropyl Ether (CAS No. 108-20-3)

A log bioconcentration factor (BCF) of 0.47 is calculated (BCF = 2.95). With respect to a log Kow = 1.52, which was used to calculate the BCF, diisopropyl ether in the aquatic environment is expected to have a low

bioaccumulation potential.

Reliability

(2) valid with restrictions

This robust summary has a reliability rating of 2 because the data are

calculated and not measured.

Reference Number

(6)

Species

other: see remark

Exposure period

at 25 °C

BCF

= 14.06

Method

other: calculation

GLP

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Remark

A log bioconcentration factor (BCF) of 1.15 is calculated (BCF = 14.06). With respect to a log Kow = 2.4, which was used to calculate the BCF,

ld 108-20-3 Date 12.16.2005

diisopropyl ether in the aquatic environment is expected to have a low

bioaccumulation potential.

Reliability

(2) valid with restrictions

This robust summary has a reliability rating of 2 because the data are calculated and not measured.

Reference Number

(6)

4.1 **ACUTE/PROLONGED TOXICITY TO FISH**

Type

flow through

Species

Pimephales promelas (Fish, fresh water)

Exposure period

96 hour(s)

Unit LC50 : mg/l = 91.7

Analytical monitoring

: yes

Method Year

other: Flow-through Fish Acute Toxicity Test

: 1983

GLP

: no data

Test substance Method

Diisopropyl Ether (CAS No. 108-20-3)

The water solubility of the test chemical was obtained from literature or determined experimentally. A flow through system using proportional diluters and modified continuous mini-diluter system was used for

maintaining the required test concentrations

Twenty to twenty-five 30 day-old fish, each weighing approximately 0.12 g, were randomly divided amongst the test tanks (control and five different

concentrations) with flow-through dilutor systems.

Lake Superior water maintained at 25°C ± 1°C was used in the test. Routine measures of hardness (EDTA) and total alkalinity of test water yielded mean values of 45.5 and 42.2 mg/L as CaCO3, respectively. The arithmetic mean of the pH was 7.5 and dissolved oxygen was always

greater than 60% of saturation.

Fish were supplied from the United States Environmental Protection Agency, Environmental Research Laboratory-Duluth culture. They were not fed during the test. Deaths were recorded after 1, 3, 6,12, 24, 48, 72, and

96 hours.

Remark

Statistics: Trimmed Spearman-Karber Method

Test method described in reference.

Result

96-hour LC50 = 91.7 mg/L based upon measured values

Analytical method used was GC analysis with Flame Ionization Detection

(GC-FID), performed on a Hewlett-Packard model 5730A gas

chromatograph. Concentrations of the test chemical were measured daily

at each exposure level.

Conclusion

96-hour LC50 = 91.7 mg/L based upon measured values.

Reliability (2) valid with restrictions

> This robust summary has a reliability rating of 2 because complete information on the analytical results were not available and the study was

not conducted under GLP.

Flag

Method

Critical study for SIDS endpoint

Reference Number

(25)

Type Calculation **Species** Fish Exposure period 96 hour(s)

Unit : mg/l LC50 = 214.1

Method other: ECOSAR version 0.99h, US EPA Test substance Diisopropyl Ether (CAS No. 108-20-3)

> ECOSAR version 0.99h, US EPA. The structure-activity relationships (SARs) presented in this program are used to predict the aquatic toxicity of chemicals based on their similarity of structure to chemicals for which the aquatic toxicity has been previously measured. Most SAR calculations in the ECOSAR Class Program are based upon the octanol/water partition

coefficient (Kow). SARs have been used by the U.S. Environmental Protection Agency since 1981 to predict the aquatic toxicity of new industrial chemicals in the absence of test data. SARs are developed for chemical classes based on measured test data that have been submitted by industry or they are developed by other sources for chemicals with similar structures, e.g., phenols. Using the measured aquatic toxicity values and estimated Kow values, regression equations can be developed for a class of chemicals. Toxicity values for new chemicals may then be calculated by inserting the estimated Kow into the regression equation and correcting the resultant value for the molecular weight of the compound.

To date, over 150 SARs have been developed for more than 50 chemical classes. These chemical classes range from the very large, e.g., neutral organics, to the very small, e.g., aromatic diazoniums. Some chemical classes have only one SAR, such as acid chlorides, for which only a fish 96-hour LC50 has been developed. The class with the greatest number of SARs is the neutral organics, which has SARs ranging from acute and chronic SARs for fish to a 14-day LC50 for earthworms in artificial soil. The ECOSAR Class Program is a computerized version of the ECOSAR analysis procedures as currently practiced by the Office of Pollution Prevention and Toxics (OPPT). It has been developed within the regulatory constraints of the Toxic Substances Control Act (TSCA). It is a pragmatic approach to SAR as opposed to a theoretical approach.

Result

LC50, 96 h, for fish = 214.1 mg/L

Test condition

Experimental water solubility, 8800 mg/l @ 20°C (Heitmann, W et al. 1987). log Kow, 1.52 (Funasaki, N et al. 1985) and melting point, -86.6°C (SRC PhysProp database) were entered into the program.

Class: Neutral organics

Conclusion

The predicted 96 h LC50 value for fish (214.1 mg/L) is in good agreement with the experimental 96 h LC50 value for fathead minnow (Pimephales promelas) (91.7 mg/L) (Veith et al., Can. J. Fish. Aquat. Sci., 40:743-748) and 48 h EC50 value for Daphnia (190.0 mg/L) (Stephenson R.R., Shell Research Limited, Report No. SBGR.83.215).

Reliability

(2) valid with restrictions

This robust summary has a reliability rating of 2 because the data are

calculated and not measured.

Reference Number

(5)

Type

flow through

Species

Pimephales promelas (Fish, fresh water)

Exposure period

96 hour(s) mq/L $= 786 \, \text{mg/L}$

Unit LC50 **EC50**

 $= 476 \, \text{mg/L}$

Analytical monitoring

Method

other: Flow-through Fish Acute Toxicity Test 1983

Year **GLP**

no data

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

conductance (78 - 86 µmhos/cm).

Method

Test solutions were prepared using a proportional diluter system without replication. This system provided control and five test substance concentrations to glass test vessels. Each vessel held 2 L of test solution and the diluter flow rate was sufficient to provide 18 volume additions per day. An aqueous stock solution of 1050 mg/L was used by the diluter to prepare the exposure series. Dilution water was filtered Lake Superior water. Typical ranges of water quality factors measured in this water were pH (7.4 - 8.2), total hardness (44 - 53 mg/L as CaCO₃), and specific

Test fish originated from in-house cultures of P. promelas at the U.S. EPA Environmental Research Laboratory – Duluth, Fish were not fed 24 h prior to testing or during the test. At test initiation, fish were randomly placed in

test vessels until each vessel contained 10 individuals. Individuals used in testing were 34 d old and measured 19.0 mm mean length (SD = 2.534) and 0.104 g mean weight (SD = 0.0433). Biomass loading was 1.04 g/L. Death was the major test endpoint. Numbers of dead fish were counted daily and any dead fish were removed from the vessels. Abnormal behavioral changes were recorded at each observation time. LC50 (lethality) and EC50 (total effect) values were determined.

Temperature, dissolved oxygen, and pH were measured daily in all test chambers. Mean values (and Standard Deviation) were 24.9 °C (SD = 0.52), 7.3 mg/L (SD = 0.13), and 7.75 (SD = 0.16), respectively. Total hardness and alkalinity were measured once in the control, low, medium, and high test levels. Mean values were 43.7 mg/L total hardness as $CaCO_3$ (SD = 0.96) and 49.6 mg/L alkalinity as $CaCO_3$ (SD = 0.25). Lighting was provided by fluorescent bulbs that produced 28 to 48 lumens/sq ft at the water surface. The photoperiod was 16 h light and 8 h dark.

Test substance concentrations were verified in most cases daily during the test using gas-liquid chromatography. Concentrations were averaged and a mean percent recovery was calculated. The nominal with measured concentrations in parentheses were, control (not detected), 157 mg/L (131 mg/L), 242 mg/L (210 mg/L), 373 mg/L (382 mg/L), 574 mg/L (594 mg/L), and 883 mg/L (1044 mg/L). The overall percent recovery was 98.7%.

Remark

Statistics: LC/EC50 values determined by Trimmed Spearman-Karber

Method

Result

: 96-hour LC50 = 786 mg/L based on mean measured values. 96-hour EC50 = 476 mg/L based on mean measured values.

The EC50 value was based on mortality and the following abnormal effects: loss of schooling behavior, swimming near the surface, hypoactive, under-reactive to external stimuli, loss of equilibrium.

Conclusion

96-hour LC50 = 786 mg/L based on mean measured values. 96-hour EC50 = 476 mg/L based on mean measured values.

Reliability

: (1) valid without restrictions

Reference Number

(7)

Type

: flow through

Species

Pimephales promelas (Fish, fresh water)

Exposure period Unit

96 hour(s) mg/L

Unit LC50

 $= 900 \, \text{mg/L}$

Limit test

Ves

Analytical monitoring Method

other: Flow-through Fish Acute Toxicity Test (ASTM, 1980)

Year GLP 1985 no data

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Method

Test solutions were prepared using a continuous-flow diluter delivery system, which delivered four test substance concentrations and control solutions to duplicate test vessels. Dilution water was filtered Lake Superior water. Average values for water quality factors for the dilution water were: hardness (44.6 mg/L as CaCO₃), total alkalinity (44.0 mg/L as CaCO₃), and pH (7.6). Test chambers were glass vessels and contained 2 L of test solution. Solution flow rates through the test chambers was sufficient to provided at least a 95% replacement in approximately 4 h. Test substance concentrations were verified daily during the test using either gas chromatography or high pressure liquid chromatography methods.

The mean temperature for the test was $25 \pm 0.5^{\circ}$ C, and dissolved oxygen remained at or above 80% saturation. Lighting was provided by wide spectrum fluorescent bulbs at an intensity of 22 to 38 lumens/sq ft over the test chambers. The photoperiod was 16 h light and 8 h dark with a 30-min

dusk/dawn transition period.

Test fish originated from cultures maintained by the U.S. EPA

Environmental Research Laboratory - Duluth, MN. and were 28 to 34 days old (weighing approximately 0.12 g) at the time of testing. A total of 20 fish per treatment (10/replicate) was used in the test. Fish were added to the test chambers 2-3 h before introduction of the test solutions. Fish were not

fed 24 h before or during the test. Mortalities were recorded daily.

Remark : Statistics: Trimmed Spearman-Karber Method or log-probit method. Result : 96-h LC50 = 900 mg/L based on measured concentrations

95% CL = 881 - 920 mg/L

Conclusion : 96-h LC50 = 900 mg/L based on measured concentrations

Reliability : (1) Valid without restrictions.

Reference Number (1.1)

Type : static

Species Carassius auratus (Fish, fresh water)

Exposure period : 24 hours Unit : mg/L LC50 = 380 mg/L

Analytical monitoring yes

Method other: static acute fish toxicity test (APHA, 1971)

Year unknown **GLP** no data

Test substance Diisopropyl Ether (CAS No. 108-20-3)

Method The test consisted of exposing groups of six fish to a series of

> concentrations of the test substance for 24 h. Fish were exposed in an all glass tanks holding 25 liters of test solution. Dilution water was local tap

water having the following characteristics (all values in mg/L):

Cl⁻ = 65; NO₂⁻ = 0; NO₃⁻ = 4; SO₄⁻² = 35; PO₄⁻³ = 0.15; HCO₃⁻ = 25; SiO₂ = 25; NH₄⁺ = 0; Fe = 0.05; Mn = 0; Ca⁺² = 100; Mg⁺² = 8; alkali as Na⁺

= 30; pH = 7.8.

The test was run at a temperature of 20±1°C, and the solutions were not aerated during the test period.

Test fish had a mean length of 6.2 ± 0.7 cm, a mean weight of 3.3 ± 1.0 g

and were in good health at the time of testing.

Exposure concentrations were confirmed either by total organic carbon analysis or by extraction and subsequent analysis by gas chromatography.

Measured concentrations were not reported in this study.

Remark Determination of LC50 by graphical interpolation of log concentrations

versus percent mortality (APHA, 1971).

Result 24-hour LC50 = 380 mg/L

> The analytical method was either total organic carbon analysis or gas chromatography. It was not reported what method was employed for this test substance nor if the result was based on measured concentrations.

Conclusion 24-hour LC50 = 380 mg/L.

Reliability (3) not reliable

> The test was run for only 24 hours to ensure that the dissolved oxygen content did not fall below 4 mg/L. The report lacked sufficient detail for assessment. It was not stated whether results were based on nominal or

measured values.

Reference Number (1)

Type static

Species Lepomis macrochirus (Fish, fresh water)

Exposure period 96 hours Unit mg/L LC50 7000 mg/L

Analytical monitoring

Method other: static acute fish toxicity test

4. Ecotoxicity

ld 108-20-3 Date 12.16.2005

Year **GLP**

not stated

Test substance

no

Diisopropyl Ether (CAS No. 108-20-3)

Method

: The test consisted of exposing groups of fish to a four-dilution series of the test substance for a period of 96 h. Test vessels were all glass 5-gallon aquaria. The volume of test solution was adjusted to assure that a biomass loading was no more than 1 g fish /liter solution. Dilution water was well water having a typical pH of 7.6 to 7.9 and a hardness of 55 mg/L (as CaCO₃).

Fish were obtained from a commercial source and assessed for health during a 14-d acclimation period prior to testing. During that time they were maintained on a commercial fish food diet supplemented with minced frozen shrimp. Fish were not fed 48 hours prior to testing. Fish were randomly selected for testing and were approximately 33 to 75 mm in length.

The test was run at 23°C. Test solutions were not aerated for the initial 24 h, but aeration was applied thereafter if the dissolved oxygen concentration was being depleted. Dissolved oxygen readings were taken daily, and pH was measured at the end of the test. However, these data were not provided in the report.

Mortality was assessed daily and any dead fish were removed at each observation time.

Remark

The LC50 was determined by plotting survival percentages on semilogarithmic paper and drawing a straight line fit through or near significant points above and below 50% survival.

Result

96-hour LC50 = 7,000 mg/L

The mortality pattern reported for the test substance suggests that a more likely estimate of the LC50 value would lie between 7,900 and 10,000 mg/L, rather than 7,000 mg/L. This was due to a non-monotonic dose response pattern of mortality. The report authors indicated that the LC50 value was higher than the published solubility for the test substance.

Conclusion Reliability

96-hour LC50 = 7,000 mg/L based on nominal concentrations

(3) not reliable

Documentation was insufficient for evaluation. Basic water quality data during the test were not provided. The authors stated that aeration of the test solutions was used after 24 hours to ensure maintenance of dissolved oxygen. No analytical verification of exposure concentrations were made. It is likely that test material was lost from the test medium during exposure.

Reference Number

(4.1)

Type Species

Exposure period

Menidia beryllina (Fish, estuarine/marine)

Unit LC₅₀ 96 hours ma/L 6600 mg/L

Analytical monitoring

Method

other: static acute fish toxicity test

Year **GLP**

not stated

Test substance

Method

Diisopropyl Ether (CAS No. 108-20-3)

The test consisted of exposing groups of fish to a four-dilution series of the test substance for a period of 96 h. Test vessels were all glass 5-gallon aquaria. The volume of test solution was adjusted to assure that a biomass loading was no more than 1 g fish /liter solution. Dilution water was prepared by adding "instant ocean" salts to well water (pH of 7.6 to 7.9; hardness 55 mg/L (as CaCO₃) until a specific gravity of 1.018 was

achieved.

Fish were field collected in nets from Horseshoe Bay at Sandy Hook, New Jersey. They were held for a 14-d acclimation period prior to testing and assessed for health during that time. During the acclimation period they

were fed minced frozen shrimp. Fish were not fed 48 hours prior to testing. Fish were randomly selected for testing and were approximately 40 to 100 mm in length.

The test was run at 20°C, and test solutions were continuously aerated during the exposure period. Dissolved oxygen readings were taken daily, and pH was measured at the end of the test. However, these data were not provided in the report.

Mortality was assessed daily and any dead fish were removed at each

observation time.

: LC50 determined by graphical interpolation of the logarithm of the

concentration versus the percentage mortality

Result : 96-hour LC50 = 6600 mg/L

The mortality pattern reported for the test substance does not correspond with the estimated LC50 value. Given the dose-response pattern, the LC50 value would lie between 3,200 and 5,000 mg/L, rather than 6,600 mg/L. The authors reported that the result was higher than the reported

water solubility of the test substance.

Conclusion Reliability

Remark

: 96-hour LC50 = 6600 mg/L based on nominal concentrations

(3) not reliable

Documentation was insufficient for evaluation. Basic water quality data during the test were not provided. The authors stated that aeration of the test solutions was used after 24 hours to ensure maintenance of dissolved oxygen. No analytical verification of exposure concentrations were made. It is likely that test material was lost from the test medium during exposure.

Reference Number

(4.1)

4.2 ACUTE TOXICITY TO AQUATIC INVERTEBRATES

Type

: Static

Species

: Daphnia magna (Crustacea)

Exposure period

48 hour(s)

Unit EC50

: mg/l : = 190

Analytical monitoring

- 10

Method

other: U.S. Environmental Protection Agency, Methods for acute toxicity testing with fish, macro-invertebrates and amphibians (EPA-660/3-75-009)

Year GLP : 1975 : No

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Remark

Statistics:

Probit analysis after log transformation of the concentrations (Finney, 1971) Probit Analysis, Finney, D.J., Cambridge University Press, 3rd edition,

p333 (1971)

Result

The 24 h and 48 h Effect Concentration (EC50) values were calculated to be 240 mg/L (95% fiducial limits 210 to 280 mg/L) and 190 mg/L (95%

fiducial limits 160 to 220 mg/L), respectively.

The immobilization (%) of Daphnia magna (n=10/replicate) are as follows:

Test Substance Immobilization (%)*
Loading Rate
(mg/L) 24 hr 48 hr

0 (control)	0	0
46	0	0
99	3	7.
210	27	57
460	100	100
1000	100	100

Test condition

*mean of 3 replicates

A 48 hour static toxicity test was carried out without renewal of the test solutions. Quantities of stock solutions of di-isopropyl ether in acetone were added in triplicate sets of 110 mL glass flasks so that when made up with reconstituted freshwater, an approximately logarithmic series of concentrations ranging from 46 to 1000 mg/L was produced. Three flasks served as controls and received no test substance. The concentration of acetone in all control and test flasks was 0.1 mL/L. Precautions were taken to (a) minimise evaporative loss of the test substance by use of glass cover slips over the vessel necks and (b) to minimize the risk of organisms becoming trapped at the surface by placing black paper caps over the flasks to create a darkened zone which the organisms would avoid. The test temperatures were in the range 20 ± 2°C, pH was in the range 8.2 to 8.4, the total hardness was 164 mg/L as CaCO3, and dissolved oxygen was in the range 8.2 to 9.2 mg/L.

The daphnids were cultured in-house, derived from a strain obtained (via ICI Brixham Laboratory) from Institut National de Recherche Chimique Applique (I.R.Ch.A.), France. Age was <24 hours old from 15 to 35 day old

All concentrations of test substance are expressed in terms of quantities initially added to the test vessels.

Conclusion

: After Daphnia magna were exposed to test solutions of di-isopropyl ether for 48 hours in a static test, the 24 h and 48 h EC50 values were calculated to be 240 mg/L and 190 mg/L, respectively.

Reliability

(2) valid with restrictions This robust summary has a reliability rating of 2 because it did not analytically verify exposure concentrations and the results are based on

nominal values.

Reference Number

(24)

Type

Calculation

Species

Daphnia (no species)

Exposure period

48 hour(s) mg/l

Unit

= 221.9

EC50 Method

other: ECOSAR version 0.99h, US EPA

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Method

ECOSAR version 0.99h, US EPA. The structure-activity relationships (SARs) presented in this program are used to predict the aquatic toxicity of chemicals based on their similarity of structure to chemicals for which the aquatic toxicity has been previously measured. Most SAR calculations in the ECOSAR Class Program are based upon the octanol/water partition coefficient (Kow). SARs have been used by the U.S. Environmental Protection Agency since 1981 to predict the aquatic toxicity of new industrial chemicals in the absence of test data. SARs are developed for chemical classes based on measured test data that have been submitted by industry or they are developed by other sources for chemicals with similar structures, e.g., phenols. Using the measured aquatic toxicity values and estimated Kow values, regression equations can be developed for a class of chemicals. Toxicity values for new chemicals may then be calculated by inserting the estimated Kow into the regression equation and correcting the resultant value for the molecular weight of the compound.

To date, over 150 SARs have been developed for more than 50 chemical classes. These chemical classes range from the very large, e.g., neutral organics, to the very small, e.g., aromatic diazoniums. Some chemical classes have only one SAR, such as acid chlorides, for which only a fish 96-hour LC50 has been developed. The class with the greatest number of SARs is the neutral organics, which has SARs ranging from acute and chronic SARs for fish to a 14-day LC50 for earthworms in artificial soil. The ECOSAR Class Program is a computerized version of the ECOSAR analysis procedures as currently practiced by the Office of Pollution

4. Ecotoxicity

ld 108-20-3 Date 12.16.2005

Prevention and Toxics (OPPT). It has been developed within the regulatory constraints of the Toxic Substances Control Act (TSCA). It is a present to SAB as appeared to a theoretical approach.

pragmatic approach to SAR as opposed to a theoretical approach.

Result

: EC50, 48 h, for Daphnia = 221.9 mg/L

Test condition : Experimental water solubility, 8800 mg/l @ 20°C (Heitmann, W et al, 1987),

log Kow, 1.52 (Funasaki, N et al, 1985) and melting point, -86.6°C (SRC

PhysProp database) were entered into the program.

Class: Neutral organics

Conclusion : The predicted 48 h LC50 value for Daphnia (221.9 mg/L) is in close

agreement with the experimental 48 h EC50 value for Daphnia (190.0

mg/L) (Stephenson R.R., Shell Research Limited, Report No.

SBGR.83.215).

Reliability : (2) valid with restrictions

This robust summary has a reliability rating of 2 because the data are

calculated and not measured.

Reference Number

(5)

4.3 TOXICITY TO AQUATIC PLANTS E.G. ALGAE

Type

: Calculation

Species

: Green Alga (not specified)

Exposure period

96 hour(s) ma/l

Unit EC50

= 134.9

ChV

: = 10.2

Method Test substance other: ECOSAR version 0.99h, US EPADiisopropyl Ether (CAS No. 108-20-3)

Method

ECOSAR version 0.99h, US EPA. The structure-activity relationships (SARs) presented in this program are used to predict the aquatic toxicity of chemicals based on their similarity of structure to chemicals for which the aquatic toxicity has been previously measured. Most SAR calculations in the ECOSAR Class Program are based upon the octanol/water partition coefficient (Kow). SARs have been used by the U.S. Environmental Protection Agency since 1981 to predict the aquatic toxicity of new industrial chemicals in the absence of test data. SARs are developed for chemical classes based on measured test data that have been submitted by industry or they are developed by other sources for chemicals with similar structures, e.g., phenols. Using the measured aquatic toxicity values and estimated Kow values, regression equations can be developed for a class of chemicals. Toxicity values for new chemicals may then be calculated by inserting the estimated Kow into the regression equation and correcting the resultant value for the molecular weight of the compound.

To date, over 150 SARs have been developed for more than 50 chemical classes. These chemical classes range from the very large, e.g., neutral organics, to the very small, e.g., aromatic diazoniums. Some chemical classes have only one SAR, such as acid chlorides, for which only a fish 96-hour LC50 has been developed. The class with the greatest number of SARs is the neutral organics, which has SARs ranging from acute and chronic SARs for fish to a 14-day LC50 for earthworms in artificial soil. The ECOSAR Class Program is a computerized version of the ECOSAR analysis procedures as currently practiced by the Office of Pollution Prevention and Toxics (OPPT). It has been developed within the regulatory constraints of the Toxic Substances Control Act (TSCA). It is a pragmatic approach to SAR as opposed to a theoretical approach.

Result

EC50, 96 h, for green algae = 134.9 mg/L ChV, 96 h, for green algae = 10.2 mg/L

Test condition

Experimental water solubility, 8800 mg/l @ 20°C (Heitmann, W et al, 1987), log Kow, 1.52 (Funasaki, N et al, 1985) and melting point, -86.6°C (SRC

PhysProp database) were entered into the program.

Conclusion

Class: Neutral organics

The predicted 96 h EC50 value for algae (134.9 mg/L) is in the same range as the predicted 48 h LC50 value for Daphnia (221.9 mg/L) and the predicted 96 h LC50 value for fish (214.1 mg/L). There is also good comparison between the predicted and experimental EC50 values for Daphnia (221.9 mg/l v 190.0 mg/L, respectively) and for fish (214.1 mg/l v 91.7 mg/L, respectively).

Reliability

(2) valid with restrictions

This robust summary has a reliability rating of 2 because the data are

calculated and not measured.

Reference Number

(5)

Type

Biomass

Species

Selenastrum capricornutum (alga, fresh water ATCC 22662)

Exposure period

96 hour(s) ma/L

Unit **EC50**

= > 1000 mg/L

Analytical monitoring

Method

Year

other: algae growth inhibition 1983

GLP

no data

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Method

A 4 d algal growth study was carried out using 10 concentrations of the test substance and a control. The test design included six control replicates and single vessels dosed with different concentrations of the test substance. 250-mL glass Erlenmeyer flasks served as the test vessels and held 50 mL of culture medium. Culture medium was prepare following the recipe given by Miller and Green (1978) with the following exceptions: 1) boric acid concentration = 105 µg/L, and 2) sodium bicarbonate concentration = 50 ma/L.

To 10 flasks, quantities of a test substance stock solution made up in acetone were added to give a logarithmic series of concentrations ranging from 1 to 1000 mg/L (1.0, 2.2, 4.6, 10, 22, 46, 100, 220, 460, and 1000 mg/L). The concentration of acetone in all flasks including controls was adjusted to 0.1 mL/L. Each flask was inoculated with S. capricornutum to give an initial cell density of 5 x 10² cells/mL. The algal inoculum was prepared from an actively growing liquid culture of S. capricornutum in exponential growth phase.

Flasks were incubated in a temperature controlled orbital incubator under constant illumination (approximately 3000 lux) at 24±2°C for 4 days. Cell counts were made on days 2 and 4 using an electronic particle counter (Coulter counter). The temperature in the incubator was measured at 4-h intervals. The pH of the control and highest test concentration was measured on days 0, 2, and 4. Temperature remained within the 24±2°C specified range, and the pH ranged from 8.3 to 8.5 in the measured vessels.

All determination of EC50 values were based on nominal test concentrations and cell counts.

Result

96-hour EC50 = >1000 mg/L based on nominal concentrations.

The 96-hour cell counts in the treated flasks as a percent of the mean control cell counts were:

1.0 mg/L = 84%46 ma/L = 127%2.2 mg/L = 108%100 mg/L = 130%4.6 mg/L = 91%220 ma/L = 113%10 mg/L = 122%460 ma/L = 127% 1000 mg/L = 91%22 mg/L = 129%

Conclusion Reliability

96-hour EC50 = >1000 mg/L based on nominal concentrations.

(3) not reliable

4. Ecotoxicity

ld 108-20-3 Date 12.16.2005

Test concentrations were not measured and there is no indication in the report whether the test vessels were sealed. The reported LC_{50} value may reflect a loss of test substance by volatilization if the flasks were not tightly sealed.

Reference Number

(24)

Result

ld 108-20-3 Date 12.16.2005

5.0 TOXICOKINETICS, METABOLISM AND DISTRIBUTION

5.1.1 ACUTE ORAL TOXICITY

Type : LD50 Species : rat

Strain : Sprague-Dawley
Sex : male/female

Vehicle : other: None; administered undiluted

Method : other: Similar to OECD 401

GLP : no

Test substance : Diisopropyl ether (CAS No. 108-20-3)

Method : Administered orally to nonfasted rats. LD50 calculated by the method of

Litchfield and Wilcoxon [1949]. Similar to OECD 401.

Remark : Test type: Acute oral toxicity

Year: Prior to 1971

No. of animals/dose: 6 male for young adult and older adult

6 - 12 male and female for 14-day old rats Route of administration: Oral gavage

Dose level: Variable Dose volume: Variable

Control group included: No, but none needed 14-day old: LD50 6.4 ml/kg [approx 4.5 g/kg] young adults: LD50 16.5 ml/kg [approx 11.6 g/kg]

Older adults: LD50 16.0 ml/kg [approx 11.2 g/kg]

G/kg dose based on a density of 0.72 g/ml

Test condition : Rats were observed for up to 7 days after dosing.

Test substance : Diisopropyl ether (CAS No. 108-20-3)

Chemical name: Isopropyl ether; isopropoxy propane, 2-; oxybis [propane],

Source/purity of test material is not specified, but stated to be analytical

grade meeting ACS specifications.

Conclusion : DIPE, when administered to adult male Sprague-Dawley rats, had an acute

oral LD50 of >10 g/kg. 14-day immature rats were considerable more

sensitive [LD50 4.5 g/kg].

Reliability : (2) valid with restrictions
Not GLP but conducted at a reputable laboratory [Abbot Laboratories,

NOT GLP but conducted at a reputable laboratory (Abbot Laboratories,

Chicago].

Reference Number (13)

Species : rabbit

Strain : New Zealand white

Sex : no data Number of animals : 6

Vehicle : other: none reported

Doses : 8.2, 7.2, 6.0, 5.2, 3.3, 1.62 g/kg **Method** : other: Similar to OECD 401

GLP : n

Test substance: Diisopropyl ether (CAS No. 108-20-3)

Remark : Test type: Acute oral toxicity

Year: Prior to 1939

Route of administration: Oral

Dose levels: 8.2, 7.2, 6.0, 5.2, 3.3, 1.62 g/kg

Dose volume: Variable Control: No - none needed

Result : Minimal lethal dose between 7 - 9 ml/kg

The symptoms noted were lack of coordination and unsteadiness at onset followed by a slight narcosis. In the animals that died the narcosis progressed towards a deep narcosis with loss of corneal reflex and evidences of depressant action on the medulla appeared, respiration became progressively slower, irregular and variable in amplitude and drop in body temperature till respiration failed. In the surviving animals, no effect on HB, erythrocyte count, total and differential leukocyte count was

observed. No delayed toxicity was observed during the recovery period of 4

months after treatment.

Test substance : Diisopropyl ether (CAS No. 108-20-3)

Chemical name: Isopropyl ether; isopropoxy propane, 2-; oxybis [propane],

2,2 '-

Test material is stated to be commercial grade with 3% of isopropyl

alcohol, but with no peroxide nor added inhibitor.

Conclusion The test article, when administered orally as received to New Zealand

white rabbits had a minimal lethal dose of 7 - 9 ml/kg [approx 5 - 6.5 g/kg].

(2) valid with restrictions Reliability

Not conducted by GLP but at a reputable laboratory [Kettering Laboratory.

University of Cincinnatil.

Reference Number

(15)

5.1.2 ACUTE INHALATION TOXICITY

Species guinea pig

Strain : other: not specified

Sex : no data Vehicle : other: none

Doses : 0.3%; 1%; 3%; 6% in air Method : other: not specified

GLP

Test substance Diisopropyl ether (CAS No. 108-20-3) Test type: Acute inhalation toxicity Remark

Year: Prior to 1939

No. animlas/sex/group: One to two animals per dose

Route of administration: Inhalation Dose level: 0.3%; 1%; 3%; 6% in air

Dose volume: N/A Control: No

Result 0.3 % (~3000 ppm)- 2 h: No visible indication of anesthetic action

1 and 3.0 % (~30000 ppm)- 1 h: Not lethal; signs of anesthesia

6.0% (~60000 ppm): Death as the result of respiratory failure within 1 hr

Test condition

1 or 2 hrs or until death [6%]

Diisopropyl ether (CAS No. 108-20-3) Test substance Chemical name: Isopropyl ether; isopropoxy propane, 2-; oxybis [propane],

2,2 '-

Test material is stated to be commercial grade with 3% of isopropyl

alcohol, but with no peroxide or added inhibitor.

Conclusion : The lowest lethal dose for 1 hr exposure was greater than 30,000 ppm in

three species.

(2) valid with restrictions Reliability

Not conducted by GLP. Few animals per group, but at a reputable

laboratory [Kettering Laboratory, University of Cincinnati].

Reference Number

(15)

Species

: rabbit

Strain : New Zealand white

Sex : no data Vehicle : other: none

Doses : 0.3%; 1%; 3%; 6% in air

ld 108-20-3 Date 12.16.2005

Method

: other: not specified

GLP

: nc

Test substance Remark Diisopropyl ether (CAS No. 108-20-3)Test type: Acute inhalation toxicity

Year: Prior to 1939

No. animlas/sex/group: One to two animals per dose

Route of administration: Inhalation Dose level: 0.3%; 1%; 3%; 6% in air

Dose volume: N/A

Control: No

Result : 0.3 % (~3000 ppm)- 2 h: No visible indication of anesthetic action

1 and 3.0 % (~30000 ppm)- 1 h: Not lethal; signs of anesthesia

6.0% (~60000 ppm): Death as the result of respiratory failure within 1 hr

Test condition

: 1 or 2 hrs or until death [6%]

Test substance

Diisopropyl ether (CAS No. 108-20-3)

Chemical name: Isopropyl ether; isopropoxy propane, 2-; oxybis [propane],

2,2 '-

Test material is stated to be commercial grade with 3% of isopropyl

alcohol, but with no peroxide or added inhibitor.

Conclusion

: The lowest lethal dose for 1 hr exposure was greater than 30,000 ppm in

three species.

Reliability

: (2) valid with restrictions

Not conducted by GLP. Few animals per group, but at a reputable

laboratory [Kettering Laboratory, University of Cincinnati].

Reference Number

(15)

Species

: monkey

Strain

other: Macacus rhesus

Sex

: female : other: none

Vehicle Doses

0.3%; 1%; 3%; 6% in air other: not specified

Method GLP

Remark

: no

Test substance

Diisopropyl ether (CAS No. 108-20-3)
Test type: Acute inhalation toxicity

Year: Prior to 1939

No. animlas/sex/group: One to two animals per dose

Route of administration: Inhalation Dose level: 0.3%; 1%; 3%; 6% in air

Dose volume: N/A

Control: No

Result

: 0.3 % (~3000 ppm)- 2 h: No visible indication of anesthetic action 1 and 3.0 % (~30000 ppm)- 1 h: Not lethal; signs of anesthesia

6.0% (~60000 ppm): Death as the result of respiratory failure within 1 hr

Test condition

: 1 or 2 hrs or until death [6%]

Test substance

Diisopropyl ether (CAS No. 108-20-3)

Chemical name: Isopropyl ether; isopropoxy propane, 2-; oxybis [propane],

2,2 '-

Test material is stated to be commercial grade with 3% of isopropyl

alcohol, but with no peroxide or added inhibitor.

Conclusion

: The lowest lethal dose for 1 hr exposure was greater than 30,000 ppm in

three species.

Reliability

: (2) valid with restrictions

Not conducted by GLP. Few animals per group, but at a reputable

laboratory [Kettering Laboratory, University of Cincinnati].

Reference Number

(15)

5.1.3 ACUTE DERMAL TOXICITY

Type

: LD50

ld 108-20-3 Date 12.16.2005

Species

: rabbit

Strain

: New Zealand white

Sex Vehicle Doses : no data : other: none : variable

Method

: other: Similar to OECD 402

GLP

. 20

Test substance

: Diisopropyl ether (CAS No. 108-20-3)

Remark

Test type: Acute dermal toxicity

Year: Prior to 1939

No. of animals/sex/group: Unspecified Route of administration: Dermal

Dose level: variable

Control: No

Result

 No deaths or systemic effects were reported. In rabbits dermal unoccluded LD50 > 2.0 g/kg. The actual dose applied was much higher, but continued

to evaporate from the skin during application.

Test condition

: The material was continuously dripped onto the shaved skin to keep it wet for one hour, while continuously evaporating. 150 ml of material was used.

Test substance

Diisopropyl ether (CAS No. 108-20-3)

Chemical Name: Isopropyl ether; isopropoxy propane, 2-; oxybis

[propane], 2,2 '-

Test material is stated to be commercial grade with 3% of isopropyl

alcohol, but with no peroxide or added inhibitor.

Conclusion

: The test article, when administered dermally to New Zealand white rabbits

had an acute dermal LD50 of greater than 2.0 g/kg.

Reliability

: (2) valid with restrictions

Not GLP but conducted at a reputable laboratory [Kettering Laboratory,

University of Cincinnati].

Reference Number

(15)

5.3 SENSITIZATION

Type

other: In vitro chemical reactivity assay, surrogate for respiratory

sensitization

Species

: other: No animals; in vitro chemical assay

Number of animals

: 0

Vehicle Result other: Nonenot sensitizingnot sensitizing

Classification

: other: No guideline available

Method Year GLP

: 1990

Test substance

: Diisopropyl ether (CAS No.108-20-3)

Remark

: Route of administration: N/A

Sex: N/A
Dose level: N/A
Dose volume: N/A

Result

Control group included: Positive and negative controls included Diisopropanol was negative in this in vitro assay for potential respiratory

sensitization. The assay gave positive responses with several known

respiratory sensitizers.

Test condition

A method for monitoring chemical reactivity in aqueous solutions, at neutral pH and 37 degrees C, was developed. The chemical was allowed to react with a lysine-containing peptide, and the reaction was monitored with high-performance liquid chromatography. Simple acids, based, and solvents did not react with the peptide, whereas isocyanates, anhydrides, and

chloramine-T, substances well known for their sensitizing and asthma inducing properties, did. Thus a positive test strongly suggested that the

ld 108-20-3 Date 12.16.2005

chemical had the potential to act as a hapten and cause sensitization when

inhaled.

Test substance Diisopropyl ether (CAS No.108-20-3)

Chemical Name: Isopropyl ether; isopropoxy propane, 2-; oxybis

[propane], 2,2 '-

Source/purity not specified.

Conclusion

Di-isopropanol was negative in this in vitro assay.

Reliability

(2) valid with restrictions

Not conducted by GLP; research method not accepted by regulatory

agencies; in vitro surrogate for respiratory sensitisation.

Reference Number

(26)

5.4 REPEATED DOSE TOXICITY

Type

Sub-chronic

Species

: rat

Sex

: male/female

Strain Route of admin. Sprague-Dawley

Exposure period

: inhalation : 6 hours/day

Frequency of treatm.

: 5 days/week for ~13 weeks

Doses

: 0, 480, 3300, or 7100 ppm

Control group

other: yes (untreated & sham-exposed) = 480 ppm

NOAEL

EPA OTS 798,2450

Method Year

GLP

1996 no data

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Method

Statistical method:

Statistical analyses of numerical data included ANOVA and Tukey's studentized range test for data on serum chemistry. Duncan's multiple range test was used for hematology and body weights to assess statistically significant differences between control and exposed groups.

Remark

Male and female rats were acclimated for 2 weeks before initiation of exposures that began at ~8 weeks of age. Exposed animals were individually housed in 1-m3 inhalation chambers. Untreated control animals were housed in a separate room in identical caging. Room environment was set to 20-22°C and 40-60% relative humidity. Lights were on a 12/12hr light/dark cycle. Food and water were provided ad libitum except during exposures.

Vapors were generated by metering DIPE on to a warmed fiberglass wick and carried to the three 1m3 exposure chambers by a stream of measured and filtered room air. Temperature and humidity in the chambers were measured every 30 minutes and air flow through each chamber was metered to provide at least 12 air changes/hour. Chamber air samples were collected with a gas-tight syringe and analyzed with a gas chromatograph with a flame ionization detector and a fused silica column. Samples were periodically obtained for analysis by gas chromatography/ mass spectroscopy.

The primary endpoints during the course of exposures were individual body weights recorded weekly and clinical signs recorded daily except on weekends.

Following the last exposure, rats were fasted overnight and weighed; blood samples were obtained via the orbital sinus using light anesthesia. Samples were used to determine values for WBC, RBC, Hgb, Hct, MCV, MCH, MCHC, platelets, and differential counts. Glucose, urea nitrogen, total protein, albumin, globulin, A/G ratio, sorbitol dehydrogenase,

aspartate aminotransferase, alanine aminotransferase, alkaline phosphatase, total bilirubin, creatinine, cholesterol, triglycerides, uric acid, Cl, Ca, Na, K, and P. Following the collection of blood samples, all animals were euthanized with sodium pentobarbital (i.p.) and exanguinated. Approximately 40 tissues were collected for histopathology and organs were weighed. Testis and associated tissues were preserved whole in 10% buffered formalin except for the left cauda epididymis of 10 rats in both control groups and the highest test group; epididymides were evaluated for morphology and number of sperm. The left testis in these groups was weighed and used for determination of number of testicular spermatids.

Type: 90-Day Subchronic

Species/strain: Rat/Sprague-Dawley; Tac:N(SD)fBR

No./sex/dose: 14/sex/group

Vehicle: None

Method: USEPA 1984, 40CFR Part 798:2450

Result

DIPE did not adversely affect clinical signs body weight, serum chemistry, hematology, or the number of sperm or spermatids. Exposure to males at 7100 ppm resulted in hypertrophy of liver cells associated with increased liver weight and in increased kidney weight with an increased incidence of hyaline droplets in proximal tubules of the kidney. Females had increased weight of both liver and kidney, although kidney increased only in relation to sham-exposed controls and no morphologic changes were observed in either organ. At 3300 ppm, weights of liver and kidney were increased in males; the liver weights were increased in females only compared to shamexposed controls and not untreated controls. No morphologic abnormalities were observed. No changes were observed with 480 ppm.

Test substance Conclusion Reliability

Diisopropyl Ether (CAS No. 108-20-3); purity ~92%.

: NOAEL = 480 ppm (2) valid with restrictions

GLP unknown. Study well documented, meets generally accepted scientific

principles, acceptable for assessment.

Reference Number

(3)

Type

: Sub-chronic : rat

Species

Sex Strain : male/female : Sprague-Dawley

Route of admin. Exposure period

: inhalation 6 hours/day

Frequency of treatm.

5 days/week for ~13 weeks 0, 450, 3250, or 7060 ppm

Control group

other: yes (sham-exposed)

Method

Doses

other: USEPA 1984, 40CFR Part 798:6050, 6400, and 6200

Year 1997 **GLP** no data

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Method

Statistical method:

All statistical analyses were performed with SAS software. Body weights, rectal temperatures fore- and hindlimb grip strengths, the number of rears, and motor activity were analyzed by a one-way analysis of variance followed by Duncan's multiple range test. The remaining data from the FOB were analyzed by Fisher's exact test using an extended contingency table containing all four groups of at given sex at a specified time. If a significant difference occurred for a given parameter, Fisher's exact test was used to directly compare each group individually against the control. Brain weights,

lengths and widths, were analyzed by Student's t-test.

Male and female rats were acclimated for 2 weeks before initiation of exposures that began at ~8 weeks of age. Sham-exposed and exposed animals were individually housed in 1-m3 inhalation chambers except during behavioral testing, when they were placed in another room overnight

Remark

and evaluated the following day. Room environment was set to 20-22°C and 40-60% relative humidity. Lights were on a 12/12-hr light/dark cycle. Food and water were provided ad libitum except during exposures. Exposure vapors were generated by metering DIPE on to a warmed fiberglass wick and carried to the three 1-m3 exposure chambers by a stream of measured and filtered room air. Temperature and humidity in the chambers were measured every 30 minutes and air flow through each chamber was metered to provide at least 12 air changes/hour. Chamber air samples were collected with a gas-tight syringe and analyzed with a gas chromatograph with a flame ionization detector and a fused silica column. Samples were periodically obtained for analysis by gas chromatography/mass spectroscopy.

Exposures were stagger-started over a 5-day period with 16 animals, 2/sex/group, receiving their first exposure on each of 5 consecutive days.

The rats were observed for signs of toxicity daily prior to initiation of exposures, and individual body weights were recorded weekly.

During weeks of Functional Observation Battery (FOB) evaluation, four rats/group would be removed from the inhalation chamber and housed separately overnight and evaluated on the following day between the hours of 7:30 a.m. and 11:00 a.m. The process was repeated for each consecutive day until all rats were evaluated. The rats were evaluated with minimal disruption to the exposure schedule and with a manageable number of rats per day, 16 on any given day. The animals were evaluated in an FOB followed by a determination of motor activity prior to initiation of exposure; the FOB following 2, 4, 8, and 13 weeks of exposures, and for motor activity following 4, 8, and 13 weeks of exposures. Following the final determination of motor activity, the animals were anesthetized, intravascularly perfused, and the brain, spinal cord, and peripheral nerves removed for microscopic examination.

The FOB consisted of initially observing home-cage positioning, posture, and reaction to removal from the cage. This was followed by evaluation for exophthalmus/palpebral closure, lacrimation, salivation, pupillary response, palpebral reflex, and pinna reflex. These observations were scored by type and intensity. The animals were then observed for open field behavior. Piloerection, respiratory rate, tremors, convulsions, posture, gait, ataxic gait, tail elevation, unperturbed activity level, vocalization, number of rears, fecal balls, and urine pools were all recorded during the open-field observations. Reactions to the approach of a pencil, finger snap, and tail pinch were ranked and recorded. Finally, fore- and hindlimb grip strength, rectal temperature, and body weight were measured. Automated motor activity was assessed for 30 minutes in figure-eight mazes after the completion of the FOB.

Following the last FOB and motor activity evaluation, the rats were anesthetized with heparinized sodium pentobarbital (i.p.). The thoracic cavity was opened and the animals were infused with phosphate-buffered gluteraldehyde through the left ventricle. The perfused brain, spinal cord, and sciatic nerve with its tibial, sural, and peroneal divisions were removed. The brain and nerve tissues were processed for embedding in paraffin or glycol methacrylate (dorsal root ganglia and peripheral nerves) and sectioned for light or electron microscopic pathologic evaluation.

Type: 90-Day Neurotoxicity

Species/strain: Rat/Sprague-Dawley; Tac:N(SD)fBR

No./sex/dose: 10/sex/group

Vehicle: None

Method: USEPA 1984, 40CFR Part 798:6050, 6400, and 6200

Motor activity in a figure-eight maze and unperturbed activity in the FOB

Result

ld 108-20-3 Date 12.16.2005

were decreased at week 4 in females exposed to 7060 ppm; activity in the FOB was also decreased in females exposed to 450 ppm at week 4. Other changes in the FOB appeared to be minor, and no changes were observed during microscopic examination of tissues from the nervous system.

Test substance Conclusion

Diisopropyl Ether (CAS No. 108-20-3); purity ~92%.

Inhalation exposures to DIPE at concentrations as high as 7060 ppm for 13

weeks resulted in few observable effects on the nervous system.

Reliability (2) valid with restrictions

GLP unknown. Study well documented, meets generally accepted scientific

principles, acceptable for assessment.

Reference Number

(21)

5.5 **GENETIC TOXICITY 'IN VITRO'**

Type

Bacterial reverse mutation assay

System of testing Test concentration

: Salmonella typhimurium : Up to 8000 ug/ml in the pre-incubation mix

Metabolic activation

: with and without

Result

negative

Method

: other: Similar to OECD Guideline 471

Year **GLP**

: 1988 : no data

Test substance

Remark

Diisopropyl ether (CAS No. 108-20-3) Strains tested: Salmonella typhimurium tester strains TA98, TA100,

TA1535, TA1537, TA1538

Exposure method: Preincubation assay for volatile compounds [Brooks and Dean 1981]

Test Substance Doses/concentration levels: Up to 8000 ug/ml in the preincubation mix

Metabolic activation: With and without (S9 fraction mix of livers of Aroclor 1254 pretreated rats)

Vehicle: Tween 80/ethanol

Tester strain, activation status, Positive Controls and concentration level: Not stated

Statistical analysis: Mean revertant colony count and standard deviation were determined for each dose point.

Dose Rangefinding Study: Cytotoxicity study

S9 Optimization Study: No

Result

: DIPE did not induce reverse gene mutation in any strain. The test substance was not genotoxic in this assay with or without metabolic

activation.

Test substance

: Diisopropyl ether (CAS No. 108-20-3)

Chemical Name: Isopropyl ether; isopropoxy propane, 2-; oxybis

[propane], 2,2 '-

Source/purity not specified.

Conclusion Reliability

: Under the conditions of this study, the test material was not mutagenic.

: (2) valid with restrictions

Restriction due to the lack of any information regarding the selection of dose levels used during the study. In addition no information is presented regarding cytotoxicity or the presence of test material precipitate in the cultures. Although study was not stated to be conducted under GLP, it was conducted at a reputable laboratory [Shell Research Limited, Sittingbourne

ld 108-20-3 Date 12.16.2005

Research Center).

Reference Number

(2)

Sister chromatid exchange assay Chinese hamster ovary cells

System of testing Test concentration

Up to 1200 ug/ml

Metabolic activation

without

Result

negative

Method

other: Similar to OECD Guideline 473

Year GLP

1984 no data

Test substance

Diisopropyl ether (CAS No. 108-20-3)

Remark

Test type: Chromosome damage

Exposure method: For volatile compounds

Metabolic activation: Metabolic activation S9 was not added because liver cells are metabolically competent

Vehicle: Tween 80/ethanol

Tester strain, activation status, Positive Controls and concentration level: Cultured CHO cells were grown in 80 cm2 flasks for 24 hr before compound treatment. Treatment periods were 5 hr in the presence of S9 mix and 24 hr in the absence of S9. Colcemid was added to all cultures 22 hr after the initial treatment. After a further 2 hr, the cells were trypsinized, resuspended in hypotonic solution and then fixed, before spotting onto slides. Cell preparations were then stained with Giemsa. The slides were randomly coded and 100 cells from each culture were analyzed microscopically. Mitotic index estimations were also made. The positive controls were ethylmethanesulfonate [-S9] and cyclophosphamide [+S9].

Vehicle control: Yes

Dose rangefinding study: Cytotoxicity study

S9 Optimization study: No

Result

DIPE did not induce chromosomal damage in CHO cells. The test substance was not genotoxic in this assay.

Test substance

Di-isopropyl ether (CAS No. 108-20-3)

Chemical Name: Isopropyl ether; isopropoxy propane, 2-; oxybis

[propanel, 2.2 '-

Source/purity of test material: 98.5%

Conclusion Reliability

: Under the conditions of this study, the test material was not mutagenic.

(2) valid with restrictions

Restriction due to the lack of any information regarding the selection of dose levels used during the study. In addition no information is presented regarding cytotoxicity or the presence of test material precipitate in the cultures. Although study was not stated to be conducted under GLP, it was conducted at a reputable laboratory [Shell Research Limited, Sittingbourne

Research Center].

Reference Number

(2)

Type

DNA damage and repair assay

System of testing Test concentration

Rat liver cells : Up to 1200 ug/ml

Metabolic activation Result

: without negative

Method Year

other: Similar to OECD Guideline 476

GLP

1984 no data

Test substance

Diisopropyl ether (CAS No. 108-20-3)

(2)

Remark

: Test type: Chromosome damage

Strains tested: RL4

Metabolic activation: Metabolic activation S9 was not added because liver cells are metabolically competent.

Vehicle: Tween 80/ethanol

Tester strain, activation status, Positive Controls and concentration level: Cultured rat liver cells were grown and treated on glass microscope slides contained in 100 ml glass Leighton tubes. After 22 hr exposure to test compound or solvent, colcemid was added to each culture. After a further 2 hr, the slides were removed, subjected to hypotonic treatment followed by fixation and stained with Giemsa. The preparations were randomly coded and 100 cells from each culture were analyzed microscopically. The positive control was 7,12-dimethylbenzanthracenene.

Vehicle control: Yes

Dose rangefinding study: Cytotoxicity study

S9 Optimization study: None needed

Result : DIPE did not induce chromosomal da

: DIPE did not induce chromosomal damage in rat liver cells. The test substance was not genotoxic in this assay.

: Di-isopropyl ether (ČAS No. 108-20-3)

Chemical name: Isopropyl ether; isopropoxy propane, 2-; oxybis [propane],

2,2 '-

Source/purity of test material: 98.5%

Conclusion Reliability

Type

Test substance

: Under the conditions of this study, the test material was not mutagenic.

(2) valid with restrictions

Restriction due to the lack of any information regarding the selection of dose levels used during the study. In addition no information is presented regarding cytotoxicity or the presence of test material precipitate in the cultures. Although study was not stated to be conducted under GLP, it was conducted at a reputable laboratory [Shell Research Limited, Sittingbourne

Research Center].

Reference Number

: Gene mutation in Saccharomyces cerevisiae

System of testing : Saccharomyces cerevisiae

Test concentration: Up to 8000 ug/ml in the pre-incubation mix

Metabolic activation : with and without

Result : negative Method : other: Similar to OECD Guideline 481

Year : 1984

GLP : no data

Test substance : Diisopropyl ether (CAS No. 108-20-3)
Remark : Test type: Yeast mitotic gene conversion

Strains tested: JD1

Exposure method: [Brooks and Dean 1981]

Metabolic activation: With and without (S9 fraction mix of livers of Aroclor

1254 pretreated rats)

Vehicle: Tween 80/ethanol

Tester strain, activation status, Positive Controls and concentration level: Yeast cells were grown in log-phase, washed and resuspended in 2/5 strength YEPD broth at a concentration of 1 X 107 cells/ml. The

suspension was divided into 1.9 ml amounts in 30 ml universal containers and 0.1 ml of test compound solution was added. For experiments with metabolic activation [+S9], 0.1 ml of DIPE was added to 01.6 ml of yeast cell suspension, together with 0.3 ml of S9 mix. Initially a range of concentrations of DIPE was tested up to 5 mg/ml if solubility allowed. A second experiment was performed based on these results and taking into account cell viability. The cultures were incubated with shaking at 30 C for 18 hr. Aliquots were plated onto the appropriate culture media for selection of mitotic gene convertants and cells surviving the treatment. Mitotic gene conversion may be scored by supplementing the minimal medium with histidine to score tryptophan prototrophs, and with tryptophan to score histidine prototrophs. Control plates were set up with solvent alone and with the positive control compounds 4-nitroquinoline oxide and cyclophosphamide.

Vehicle control: Yes

Dose rangefinding study: Cytotoxicity study

\$9 Optimization study: No

DIPE did not induce mitotic gene conversion I yeast. The test substance

was not genotoxic in this assay with or without metabolic activation.

Test substance Di-isopropyl ether (CAS No. 108-20-3)

Chemical name: Isopropyl ether; isopropoxy propane, 2-; oxybis [propane],

Source/purity of test material: 98.5%

Conclusion

Result

: Under the conditions of this study, the test material was not genotoxic. Reliability

: (2) valid with restrictions

Restriction due to the lack of any information regarding the selection of dose levels used during the study. In addition no information is presented regarding cytotoxicity or the presence of test material precipitate in the cultures. Although study was not stated to be conducted under GLP, it was conducted at a reputable laboratory [Shell Research Limited, Sittingbourne

Research Center].

Reference Number

(2)

5.8.2 DEVELOPMENTAL TOXICITY/TERATOGENICITY

Species

: rat

Sex

female

Strain

Sprague-Dawley inhalation

Route of admin. Exposure period

6 hr/day

Frequency of treatm.

Gestation Days 6-15

Duration of test

20 days

Doses

0, 430, 3095, or 6745 ppm

Control group

other: yes (untreated & sham-exposed)

other: NOEL Maternal other: NOEL Pup

= 430 ppm

= 430 - ppm

Result

Maternal NOEL: 430 ppm; Pup NOEL: 430 ppm

Method

EPA OTS 798.4350

Year **GLP**

1996 no data

Test substance

Diisopropyl Ether (CAS No. 108-20-3)

Method

Statistical method:

Statistical analyses of numerical data included ANOVA and Tukey's studentized range test for data on serum chemistry. Duncan's multiple range test was used for hematology and body weights to assess statistically significant differences between control and exposed groups.

Remark

Data on the maternal biophase, cesarean sections, and fetuses were evaluated by ANOVA followed by group comparisons using Fisher's exact or Dunnett's test.

Nulliparous females were housed with males in a 1:1 ratio and observed daily for evidence of breeding activity. Females positive for sperm plug and for sperm in the vaginal lavage fluid were considered to be at day 0 of gestation and were individually housed. The females were then randomly distributed to 5 groups of 22 animals each: untreated controls, shamexposed controls, and 3 groups exposed to vapors of DIDP at 430, 3095, or 6745 ppm for 6 hr/day on gestation days (GD) 6-15.

Vapors were generated by metering DIPE on to a warmed fiberglass wick and carried to the three 1m3 exposure chambers by a stream of measured and filtered room air. Temperature and humidity in the chambers were measured every 30 minutes and air flow through each chamber was metered to provide at least 12 air changes/hour. Chamber air samples were collected with a gas-tight syringe and analyzed with a gas chromatograph with a flame ionization detector and a fused silica column. Samples were periodically obtained for analysis by gas chromatography/ mass spectroscopy.

Sham-exposed and test animals were housed in their exposure chambers throughout the exposure period; untreated controls in a separate room. Food and water were not available during the 6 -hour exposure periods but were available ad libitum at all other times. All animals were observed daily. Body weights were recorded on days 0, 6, 13, 16, and 20. Food consumption was measured on GD 6, 13, 16, and 20. Females were sacrificed on GD 20 by diethyl ether overexposure followed by exsanguination. Serum samples from the descending aorta were analyzed for glucose, urea nitrogen, total protein, albumin, globulin, A/G ratio, sorbitol dehydrogenase, lactate dehydrogenase, aspartate aminotransferase, alanine aminotransferase, alkaline phosphatase, total bilirubin, creatinine, cholesterol, triglycerides, uric acid, Cl, Ca, Na, K, and P. All organs were examined grossly. The number of corpora lutea per ovary and the gravid uterine weights were recorded. Uterine contents were examined and the numbers of implantation sites, early and late resorptions, and live and dead fetuses were recorded. The gender of each fetus was recorded. Fetuses were weighed and examined for gross anomalies. Fetuses of each litter were equally distributed between two groups; half were fixed in Bouin's solution and examined for visceral anomalies and the remaining fetuses were fixed in 95% ethanol and examined for skeletal anomalies after differential staining for cartilage and bone.

Dams exposed to 6745 ppm had a slight reduction in body weight gain and a significant decrease in food consumption. A concentration-related increase in the incidence of rudimentary 14th ribs was observed (statistically significant at 3095 and 6745 ppm) but the relevance of the finding was uncertain. There was no apparent toxicity, either maternal or fetal, at the lowest exposure concentration, 430 ppm.

Type: Developmental Toxicity

Species/strain: Sprague-Dawley; VAF/Plus Crl:CD(SD)BR

No./dose: 22/group Vehicle: None

Method: USEPA 1984; 40CFR Part 798:4350

Maternal NOEL: 430 ppm Pup NOEL: 430 ppm

Diisopropyl Ether (CAS No. 108-20-3); purity ~92%.

DIPE is not a teratogen.

(2) valid with restrictions

GLP unknown. Study well documented, meets generally accepted scientific principles, acceptable for assessment.

Result

Test substance Conclusion Reliability

Reference Number (3)

5.11 ADDITIONAL REMARKS

Type

other: Sensory Irritation in Humans

Method

Non-guideline.

Remark

Species/strain: Humans Sex: Male and female

Number/sex/group: Average of 12 Route of administration: Inhalation

Vehicle: None Control: No Year: Prior to 1946

GLP: No

Result

300 ppm: 35% of the subjects objected to this solvent because of the

unpleasant odor rather than irritation.

500 ppm: there was a sensory response that was acceptable to the

majority of subjects.

Test condition

: Subjects were exposed for 15 minutes and olfactory fatigue and irritation of mucous membranes were reported. "Motion pictures were shown to occupy the subject's attention and divert their thoughts from the atmospheric

contamination to which they were exposed."

Test substance

: Diisopropyl ether (CAS No. 108-20-3)

Chemical Name: Isopropyl ether; isopropoxy propane, 2-; oxybis

[propane], 2,2 '-

Test material is stated to be technical grade product.

Conclusion

: DIPE does not appear to be a sensory irritant at concentrations up to 500

ppm, but it does have an unpleasant odor at this concentration.

Reliability

(2) valid with restrictions

Not GLP but conducted at a reputable laboratory [Harvard School of Public

Health, Boston].

Reference Number

(23)

Type Method other: Sensory irritation in humans

Non-Guideline.

Remark

Species/strain: Young adult humans [University of California staff and

medical students]
Sex: Not specified

Number/sex/group: Not specified Route of administration: Inhalation

Vehicle: None Control: No Year: 1955 GLP: No

Result

Numbers of subjects with degree of effect

Concentration 400 ppm 800 ppm

Number subjects: 7

Eye irritation: 7 absent 3 absent, 3 slight, 1 mod.

Nose irritation: 5 absent, 2 slight 2 absent, 5 slight

Pulmonary discomfort: 7 absent 4 absent, 3 slight

Olfactory cognition: 1 slight, 6 mod. 4 mod., 3 severe

CNS effects: 7

7 absent 7 absent

Test condition

Exposures were conducted in a whole-body chamber approximately 7700 l

equipped with a fan. Exposures were made in a static atmosphere

generated by vaporizing a predetermined quantity of test solvent from a hot surface. Five minutes were allowed for evaporation and equilibration, and subjects were exposed for 5 minutes, during which time they noted the

degree of subjective responses at one-minute intervals.

Test substance

Diisopropyl ether (CAS No. 108-20-3)

ld 108-20-3 Date 12.16.2005

Chemical Name: Isopropyl ether; isopropoxy propane, 2-; oxybis

[propane], 2,2 '-

Test material is stated to be commercial grade with purity of 98% or better,

provided by Shell Chemical Corporation.

Conclusion

400 ppm: 5 mins of inhalation exposure caused no eye irritation, none to slight nose irritation, no pulmonary discomfort, olfactory recognition but no central nervous system effects.

800 ppm: 5 mins of inhalation exposed caused slight eye and nose irritation, none to slight pulmonary discomfort, definite olfactory recognition but no central nervous system effects.

Reliability

: (2) valid with restrictions

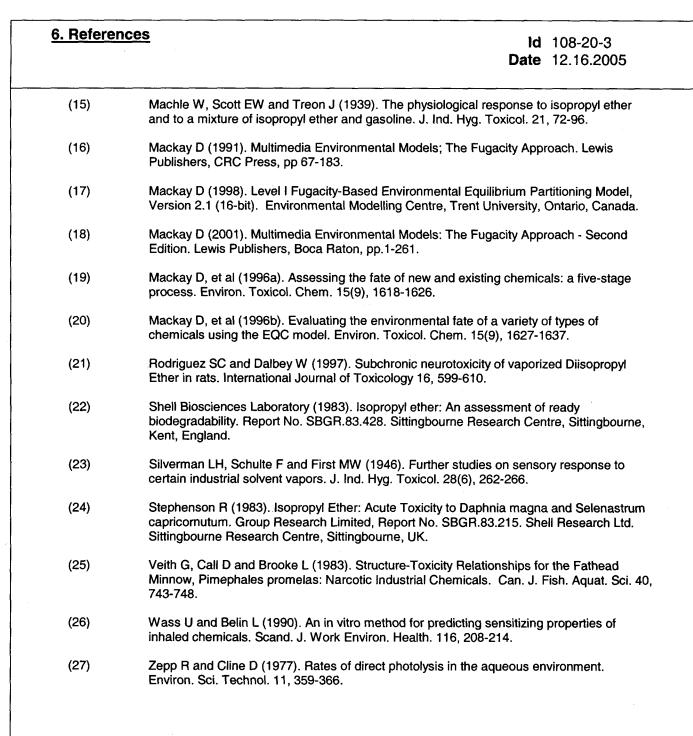
Not GLP but conducted at a reputable laboratory [University of California

School of Medicine].

Reference Number

(12)

Bridie A, Wolff C and Winter M (1979). The acute toxicity of some petrochemicals to (1)goldfish, Water Res. 13, 623-626. (1.1)Broderius S and Kahl M (1985). Acute toxicity of organic chemical mixtures to the fathead minnow. Aquat. Toxicol. 6, 307-322. (2)Brooks T, Meyer A and Hutson D (1988). The genetic toxicology of some hydrocarbon and oxygenated solvents. Mutagenesis 3(3), 227-232 Dalbey W and Fueston M (1996). Subchronic and Developmental Toxicity Studies of (3)Vaporized Diisopropyl Ether in Rats. Journal of Toxicology and Environmental Health 49, 29-43. Daubert T and Danner R (1989), Physical and thermodynamic properties of pure (4)chemicals: Data compilation, Design Institute for Physical Property Data, American Institute of Chemical Engineers. Hemisphere Publishing Corp., New York, NY, USA. (4.1)Dawson G, Jennings A, Drozdowski D and Rider E (1975/77). The acute toxicity of 47 industrial chemicals to fresh and saltwater fishes. J. Haz. Mat. 1, 303-318. Eadsforth C (1983). Isopropyl ether: determination of the n-octanol/water partition (4.2)coefficient using a reverse-phase HPLC method. Report # SBGR.83.131. Shell Research Limited, Sittingbourne Research Centre, Sittingbourne, Kent, England. ECOSAR v0.99h in EPI Suite (2000). Estimation Program Interface Suite, v3.12. Syracuse (5) Research Corporation, Syracuse, NY, USA. EPI SuiteTM (2000). Estimation Program Interface Suite, v3.12. Syracuse Research (6) Corporation, Syracuse, NY, USA. Geiger D, Poirier S, Brooke L and Call D (eds.) (1986). Acute Toxicities of Organic (7)Chemicals to Fathead Minnows (Pimephales promelas). Vol. 3. Center for Lake Superior Environmental Studies, Univ. of Wisconsin-Superior, Superior, WI, USA. (8) Gerhartz W, Yamamoto Y, Kaudy L, Rounsaville J and Schulz G (eds.) (1987). Ullmann's Encyclopedia of Industrial Chemistry. Vol. A 10. 5th Edition. VCH Publishers, New York, NY, USA. (9) Gould E (1959). Mechanism and Structure in Organic Chemistry. Holt, Reinhart and Winston, New York, NY, USA. (10)Hansch C, Leo A and Hoekman D (1995). Exploring QSAR - Hydrophobic, Electronic and Steric Constants, p. 6. ACS Professional Reference Book, American Chemical Society, Washington, DC, USA. Harris J (1982). Rate of Hydrolysis. In: Handbook of Chemical Property Estimation (11)Methods. Chapter 7. Edited by WJ Lyman, WF Reehl and DH Rosenblatt, McGraw-Hill Book Company, New York, NY, USA, (12)Hine CH, Anderson HH and Kodama JK (1955), Sensory thresholds of certain Shell organic solvents, Progress Report 1, Report to Shell Development Company, November 15, 1955. UC Report No. 247. (13)Kimura ET, Ebert DM and Doge PW (1971). Acute toxicity and limits of solvent residues for sixteen organic solvents. Toxicol. Appl. Pharmacol. 19, 699-704. Lide D, et al. (eds.) (1997-1998). CRC Handbook of Chemistry and Physics. 78th Edition. (14)CRC Press, New York, NY, USA.



International Uniform Chemical Information Database

Information on International Uniform Chemical Information Database (IUCLID) can be found on the European Chemicals Bureau (ECB) website (http://ecb.jrc.it/) by searching under the topic "IUCLID". The ECB website provides data and information on the chemical assessment procedure used by the European Union.